

An evaluation of the phytotoxicity of municipal solid waste during co-composting with different animal manures

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Abstract

Composting is a biooxidative process carried out under controlled conditions which utilizes naturally occurring microorganisms for organic waste recycling. In this study, a 60 days co-composting experiment of different treatment consisting of municipal solid wastes and plant wastes with cow dung, poultry dropping and swine manure were investigated. Compost maturity parameters and phytotoxicity, during co-composting of municipal solid waste (MSW) with different animal manures (3:1 w/w) were evaluated. Four composting treatment set-ups: MSW (control), MSW+Swine Manure (SM), MSW+Poultry Manure (PM) and MSW+Cattle Manure (CM) were prepared and composting carried out for 60 days. Changes in temperature and pH were measured daily while Electrical Conductivity, Organic Matter, Total Organic Carbon, Total Kjeldahl Nitrogen, and carbon/nitrogen ratios were assessed at day 0 and day 60. Humification process and phytotoxicity of the treatments were assessed at days 0, 10, 20, 30, 40, 50 and 60, using maize seeds germination bioassays. MSW+PM treatment showed the highest temperature of 63°C during the thermophilic phase while the control showed the least. The pH at the end of composting in all treatments exhibited alkaline values with MSW+CM showing the highest value of 8.6. Maturity indices showed that there were significant decreases in organic matter, C/N and NH₄⁺ -N/NO₃⁻ -N ratios in all the treatments when compared with the control. At the end of composting, Germination index (GI) values varied from 52.56 to 97.23% with MSW+PM showing the highest germination index (GI) of 97.23±6.08% followed by MSW+CM and MSW+SM with GI values of 88.4±6.3 and 84.8±5.3% respectively. The germination index indicated that there were significant reductions in the phytotoxicity of the treatments.

Keywords: Composting, municipal solid waste, phytotoxicity, poultry manure, seed germination test

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Introduction

Global concerns on the environment in recent times have encouraged world leaders to put in place legislative and regulatory measures aimed at promoting sustainable development in many countries worldwide. One key feature of these measures has been the greater emphasis on reducing waste volume through recycling as well as a move away from solid waste landfilling (Cunha-Queda, 2007; Xiao et al., 2009). Solid waste management through effective disposal methods has become a considerable issue albeit a challenging task to the relevant government agencies especially

for densely populated cities in many developing countries.

In Nigeria, the volume of municipal waste and bio solids has ever been increasing, particularly in the major cities and towns, owing ostensibly to increase in population and by extension anthropogenic activities (Ogunwande et al., 2008). Presently, the main municipal solid waste (MSW) disposal method is through unsanitary landfilling and incineration at open dump sites. This practise produces much of secondary pollutants, such as landfill leachate, greenhouse gases and odour.

Presumably, owing to the enormous

investment, long processing cycle and unstable products in a conventional composting treatment, only a few MSW are treated by composting in Nigeria. Composting, whose chief aim is to convert organic waste into a relatively stable material reduces the volume and stabilizes the waste to obtain a final product that is stable, free of pathogens and without phytotoxicity to plants (Choy et al., 2015; Selim, 2012; Nas and Bayram, 2008). Consequently, a composting process must be suitably managed and the progressive changes carefully controlled to give an end-product with optimum qualities which is often characterised in terms of its maturation and stability (Vergnoux et al., 2008).

In the evaluation of compost quality, a number of physico-chemical parameters such as pH, electrical conductivity, carbon form, inorganic nitrogen forms and cation exchange capacity and biological parameters such as germination index have been used successfully as indicators of compost stability and maturity of different compost mixes. In this regard, one of the most widely accepted parameters is the seed germination test which is also a measure of the phytotoxicity (CunhaQueda et al., 2007).

Thus the purpose of this investigation is to comparatively evaluate some maturity indexes and phytotoxicity of municipal solid waste co-composted with different manures in a rotary drum (composter) under 60 days, using the germination tests of maize seeds (*Zea mays*).

Materials and methods

Composting and sampling

The municipal solid waste (MSW) was collected from dump site in Ile-Ife, Osun state, Nigeria. It was source selected before use. Fresh cow dung was obtained from the Adesanmi abattoir, Ede road, Ile-Ife while the poultry manure and swine manure were obtained from the respective animal units at the Teaching and Research Farm of the Obafemi Awolowo University. The experimental set-ups were made up of four composting treatment mixes consisting of thirty kilograms (30 kg) each of a mixture of the MSW and animal manures (3:1, w/w) as follows: (i) MSW (Control) (ii) MSW + Poultry Manure (PM) (iii) MSW + cattle manure (CM) and MSW + Swine manure (SM). Composting was carried out in a perforated rotary drum composter (H = 0.35m, D = 0.30m) for 60 days and the drums were rotated once every 3 days. Moisture content was maintained to 60-65% during composting. Triplicate compost samples were collected from the rotary drum, air-dried, ground to below 0.25 mm

and subjected to analysis.

Physico-chemical analysis

The moisture and organic matter (OM) contents of the samples were determined after drying at 105°C for 24 h and ashing at 55°C for 4 h respectively (Curtin et al., 2012). The pH and electrical conductivity (EC) were determined in aqueous extracts (1:10) using a Hanna Digital Combo Meter (H1991405, Hanna, UK). The NO_3^- -N and NH_4^+ -N were determined in 0.5 M K_2SO_4 extracts (1: 10, w/v) by the modified indophenol blue technique (Barrena et al., 2014), with a microplate reader (Bio-Rad Model 550). Total N (TKN) was by semi-micro Kjeldahl analysis. Available phosphorus was determined in extracts of 0.5 M NaHCO_3 (1:100) followed by analysis with the molybdate-ascorbate method while the total organic carbon (TOC) was determined according to the colorimetry method as described (Gabhane et al., 2012). The C/N ratio was computed from contents of Total Organic Carbon (TOC) and total nitrogen in dried samples. The humic acid and fulvic acid concentrations were determined on the test sample extracts using 0.1 M $\text{Na}_4\text{P}_2\text{O}_7$ + NaOH extraction and precipitation at pH 2 followed by spectrophotometric measurement at 465 and 665 nm respectively (Huang et al., 2015).

Germination test

The seed germination tests were performed with maize seeds (*Zea mays*) according to the method described by Selim et al., 2012. The test were carried out on the 10% crude aqueous extracts of the compost sample. A quantity (0.5 ml) of each extract was pipetted into a sterilized plastic petri dish lined with a Whatman No 2 filter paper. Ten maize seeds were then evenly placed on the filter paper and incubated at 25°C in the dark for 72 hrs after which treatments were evaluated by counting the number of germinated seeds and measuring the length of roots. Three replicates for each sample of the compost. The germination index (GI) was computed using the following formula:

$$G.I = (Gt \times Lt) / (Gc \times Lc) \times 100$$

Where: Gt = % of seed germination in test sample

Gc = % of seed germination in control

Lt = root length in test sample

Lc = root length in control

(Tiqua, 2010; Selim et al., 2012)

Statistical analysis

Data are expressed as means \pm Standard error of the mean (SEM) of three independent sample replicates. One way analysis of variance (ANOVA) was performed to compare the means of treatment set-ups. Difference between means were statistically analysed using a software package, GraphPad® InStat (version 1.4). The least significant difference test at $p=0.05$ was

carried out to compare means.

Results and Discussion

The four composting treatment set-ups achieved thermophilic temperatures (>45°C) after 10 days (Fig. 1). These temperatures were maintained for 15–25 days, and gradually declined to ambient values thereafter. The thermophilic phase lasted for 10 days longer in MSW control than in the other treatment set ups. The duration of high temperatures is consistent with results from similar studies, and has been reported to reflect the usually high proportion of degradable substances that are found in this type of waste (Tognetti et al., 2008, Curtin et al., 2012 and Sundberg et al., 2013). Though these temperatures do not correspond to optimum biological activity (around 40°C), some other studies however have reported composting temperatures as high as 60–70°C during municipal organic waste composting which has been attributed to a consequence of heat accumulation in the treatment piles rather than corresponding to optimum biological activity of the microorganisms

(Ko et al., 2008). It has been reported that an excessive duration of thermophilic phase (beyond 6 weeks) as observed in the MSW (control) treatment may indicate an abnormally extended decomposition and a delayed transition to the stabilization stage (Satisha and Devarajan, 2007).

The pH values observed were in the range between 6.3 and 8.6 (Fig. 2). The lowest pH values (6.3) at the start of composting was observed in the MSW + SM treatment while the highest 8.6 was observed in MSW + PM. No consistent pH trend was observed for all the treatment set ups throughout the duration of the composting experiment, however, the pH values in all treatments increased to the alkaline range at the end of composting with MSW + PM showing the highest of 8.3. Electrical conductivity (EC), values which varied between 1.53 and 2.16 dS/m (Table 1), increased in all treatments from day 0 to day 60 (Fig. 3). The values of pH and EC achieved at 60 days of composting were within the range acceptable for plant growth as recommended by Said-Pullicino et al., 2007 and similar to other values reported for Municipal organic waste (MOW) composts (Tognetti et al., 2008).

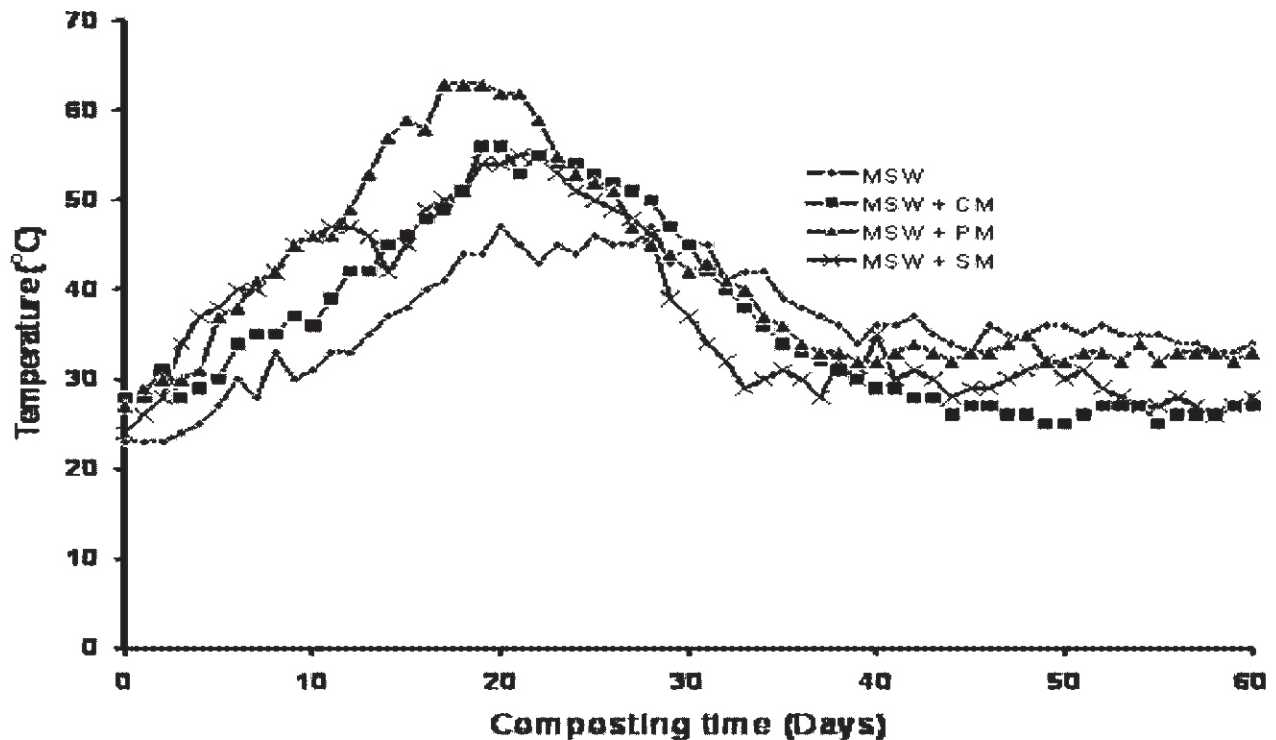


Figure 1: Changes in temperature during the composting of municipal solid waste with different animal manure

Values represent mean ± SEM, n=3

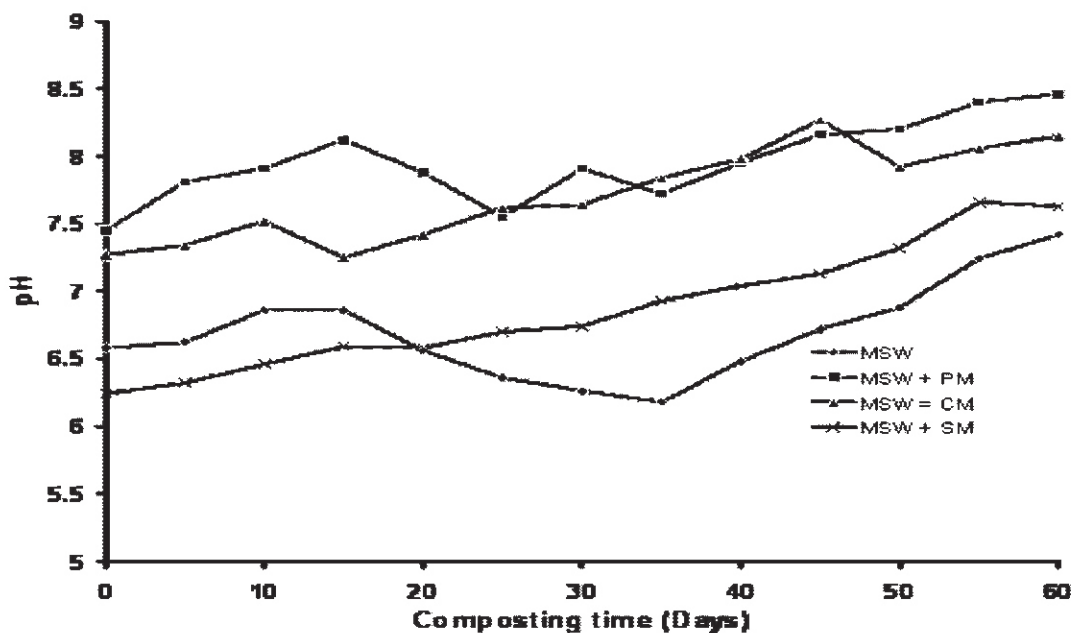


Figure 2: Changes in pH during the composting of municipal solid waste with different manures. Values represent mean \pm SEM, n=3

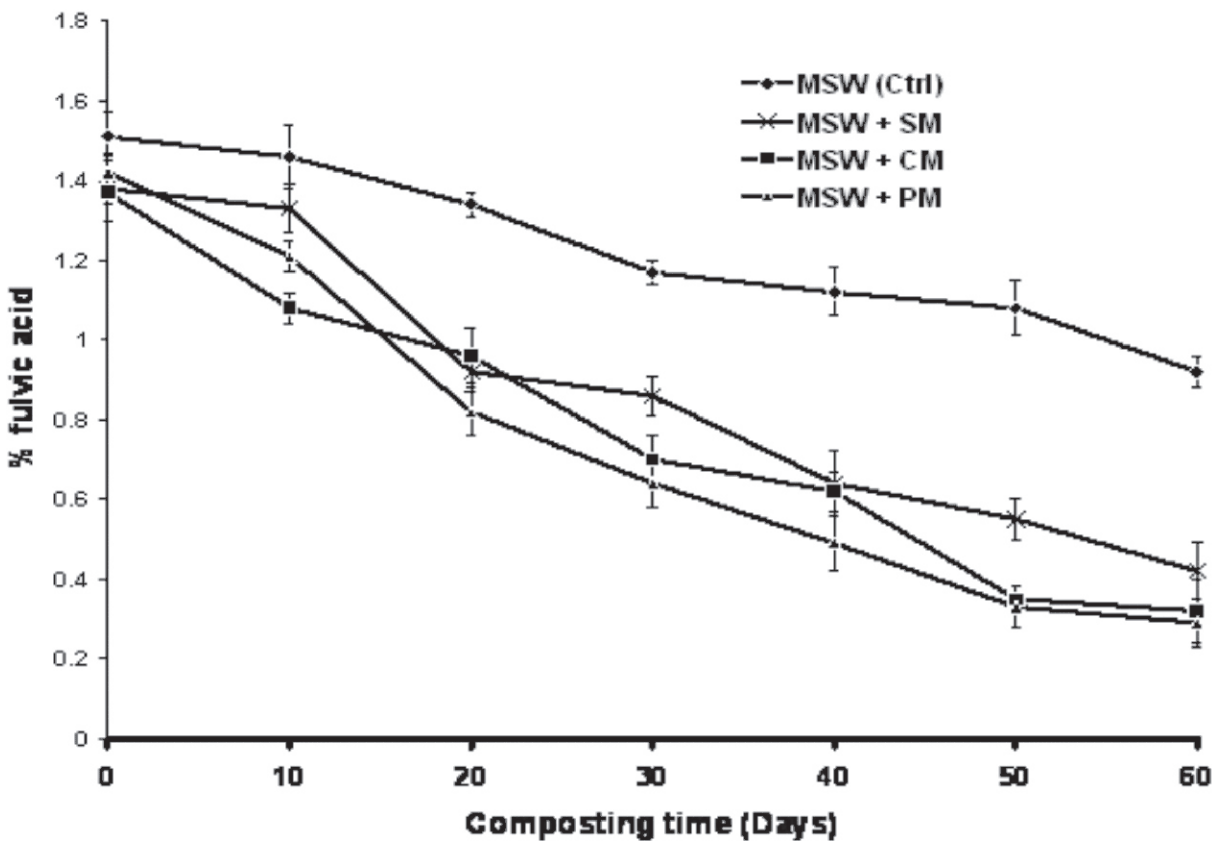


Figure 3: Changes in % fulvic acid during the composting of municipal solid waste with different manures.

Values represent mean \pm SEM, n=3, * significantly different from MSW (Ctrl), (p < 0.05)

Table 1. Physicochemical parameters at the initial and final stages composting of municipal solid waste with different animal manures

PARAMETERS	MSW		MSW + CM		MSW + SM		MSW + PM	
	Initial	Final	Initial	Final	Initial	Final	Initial	Final
EC (dSm ⁻¹)	1.14 ±	1.53 ± 0.08†	1.26 ± 0.08	1.97 ±0.11†	1.3 ± 0.12	1.88 ±0.10†	1.35 ± 0.07	2.16 ± 0.12†
MOISTURE (%)	15.1 ± 3.5	34.7 ± 4.7	35.8 ± 2.7	40.2 ± 4.2	34.7 ±	41.6 ± 2.5	66.4 ± 5.4	73.6 ± 6.3
ASH	53.5 ± 4.8	60.4 ± 5.5	53.6 ± 3.7	71.1 ± 3.2	42.8 ± 5.2	79.2 ± 7.1	56.2 ± 6.2	73.6± 6.2
OM (%)	87.2 ± 2.8	54.6 ± 1.7	76.4 ± 3.2	46.8 ± 2.5	82.6 ± 3.8	48.1 ± 2.6	83.9 ± 4.7	44.7 ± 3.1
TOC (g/Kg dw)	419.5 ± 13.6	226.5 ± 14.3	357.6 ± 23.2	195.7 ± 7.2†	345.7 ± 15.9	206.8 ± 14.1†	388.6 ± 25.4	204.2 ± 16.3†
TKN (g/Kg dw)	9.2 ± 2.5	20.4 ± 1.3	23.4 ± 2.5	30.7 ± 4.2	24.1 ± 2.5	34.2 ± 3.6	27.3 ± 4.2	34.8 ± 3.6
C/N	31.2 ± 6.5	25.54 ± 3.6	26.7 ± 2.5	17.6 ± 3.3*	31.3 ± 2.8	19.4 ± 1.7*	28.8 ± 2.1	15.7 ± 2.1†*
NH ₄ ⁺ -N (mg/Kg dw)	410.8 ± 7.8	316.3 ± 20.7	534.9 ± 10.5	245.8 ± 13.7	463.6 ± 37.2	202.4 ± 16.4	394.6 ± 25.6	216.5 ± 15.5
NO ₃ ⁻ -N (mg/Kg dw)	26.8 ± 4.6	105.4 ± 8.5	78.6 ± 7.3	134.5 ± 16.4	56.8 ± 10.2	176.5 ± 23.8	45.4 ± 5.8	184.5 ± 14.2
NH ₄ ⁺ / NO ₃ ⁻	15.3	3.1	6.8	1.8	8.2	1.1	8.7	1.2
Phosphorus (mg/Kg dw)	312.1 ± 12.3	176.8 ± 31.8	267.5 ± 12.1	322.7 ± 10.6†*	159.6 ± 17.5	245.7 ± 18.3†*	201.3 ± 24.2	389.7 ± 23.8†*

Each data represent mean ± SEM, n=3, * significantly different from final MSW (Ctrl), † significantly different from initial value. (p<0.05)

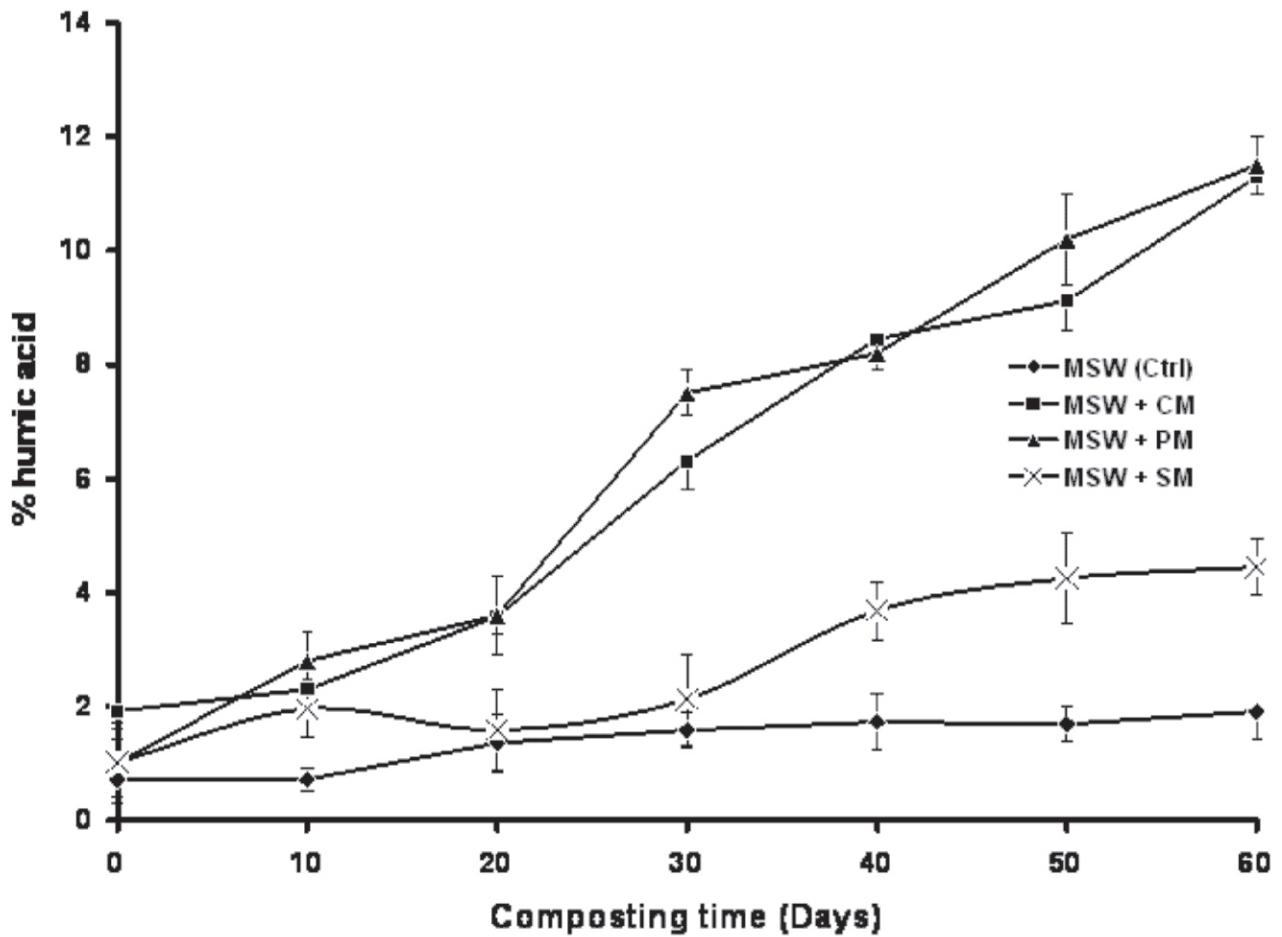


Figure 4. Changes in % humic acid during the composting of municipal solid waste with different animal manures

Values represent mean ± SEM, n=3, * significantly different from MSW (Ctrl), (p < 0.05)

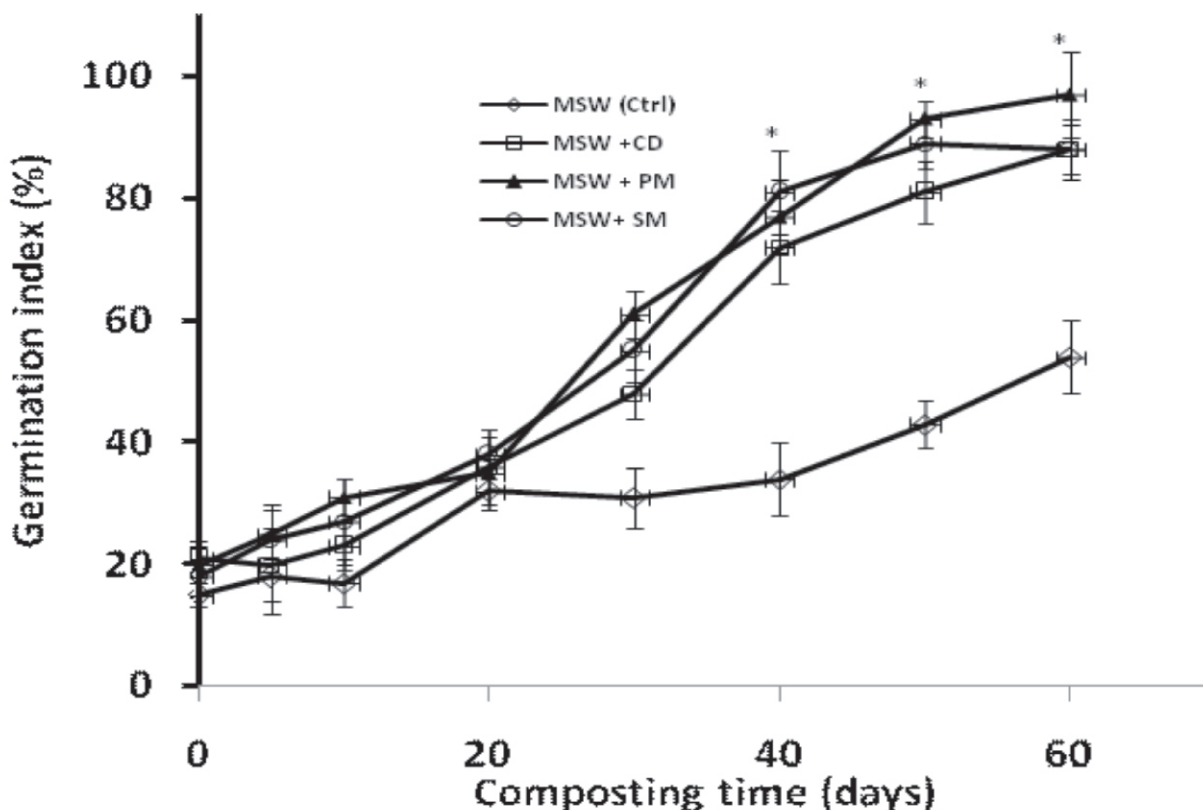


Figure 5: Changes in germination index during the composting of municipal solid waste with different animal manures

Values represent mean \pm SEM, n=3, * significantly different from MSW (Ctrl), ($p < 0.05$)

Other studies have attributed alkaline values of municipal organic waste composts to the presence of wood ashes in the wastes (Tognetti et al., 2008), which could be a common occurrence in the area from where the MSW samples used in this study was obtained, being a semi urban city. Kong, D. et al., (2012) had earlier noted that the change of mesophilic to thermophilic conditions during municipal waste composting coincided with a change in pH from acidic (<6) to alkaline (8–9). These alkaline values were attributed to proton consumption during decomposition of volatile fatty acids, which is usually abundant in the initial mesophilic phase, during which organic-N mineralization to NH_4^+ -N occurs as a result of intensive proteolysis during protein degradation (Satisa and Devarajan, 2007).

In the present study, MSW + PM had the highest pH values, possibly due to intensive thermophilic phase, which could have favoured extensive organic-N ammonification when compared to the other treatment set-ups. On the other hand there were no significant differences between the final EC values of the different compost treatments, however, EC value increased significantly ($p < 0.05$) between initial values and final values in all the treatment set-ups including

the control (Table 1). A sharp increase in EC values due to the release of soluble salts like ammonium and phosphate after the degradation of the most labile compounds in the thermophilic stage of composting has been reported by several investigators (Tiquia 2010). It is well established, that EC reflects the salinity of an organic amendment and that high salt concentration has been reported to cause phytotoxicity problems and therefore EC is a good indicator of the suitability and safety of compost for agricultural purposes (Tiquia 2010). Since none of the treatment set-up had EC value exceeding the threshold value of 4 dS m^{-1} it may indicate that the compost samples can be safely applied to soil (Ko et al., 2010).

With respect to the maturity and stability parameters measured, the results showed that at the end of the composting period most of the observed values had attained thresholds as recommended by different investigators (Table 1). For instance, in all the treatments except for the MSW (Ctrl) the OM $< 10 \text{ mg/kg}$, C/N < 20 , NH_4^+ -N $< 400 \text{ mg/kg}$, NO_3^- -N $> 160 \text{ mg/Kg}$ NH_4^+ -N/ NO_3^- $< 68 \text{ mg/Kg}$ (Bernal et al., 2009, Raut et al., 2008; Choy, 2015.) High rates of OM degradation, and substrate transformation are characteristic of active composting (Saaglia et al., 2011). Decreases of OM and C/N during the

composting process have been widely reported (Barena et al., 2008; Antil, 2014), to be indicative of active mineralization of OM by microorganisms. This observation apparently indicates all the three manures stock are suitable for composting of municipal organic waste and that maturity of compost could be achieved within a 60 days period when carried out under controlled composting conditions as employed in this study such as the use of a rotary drum.

As shown in Fig. 4, the concentration of humic acid as a percentage of humic substance increased gradually over the first 20 days and then increased sharply from 25 to 60 days in all the treatment groups except in MSW (ctrl). MSW + PM and MSW + CM showed quite a remarkable and significant ($p < 0.05$) increases compared to MSW + SM. By contrast, the concentration of fulvic acid (FA) decreased steadily in all the treatment set-ups during the composting process. This observation is consistent with that obtained in other similar investigations which related increased level of HA or decrease in FA with the occurrence of humification process. Several studies have used humification parameters as index for compost maturity (Antil, 2014). He also reported that the CHA/CFA ratio greater 1.6 is indicative of compost maturity.

Selim et al., (2012) using a combination of gel chromatography, E_4/E_6 ratio, FTIR and fluorescence spectroscopy reported the formation of polycondensed higher molecular weight compounds through aromatization to account for the increases in HA as an indicator of humification and maturity in the co-composting of pig manure and sawdust.

Fig. 5 shows the changes in GI of the different treatment set-ups during 60 days of composting. The GI increased gradually in all treatment set ups during first 30 days of composting indicating that the elimination of phytotoxicity was slow. However, from day 21 onwards, the GI in all the treatment set up increased sharply and were significantly different when compared to the MSW (Ctrl) from day 40. The germination index which is a relative seed germination ratio to that of root growth is an indicator of phytotoxicity and one of the veritable index in assessing compost maturity and stability (Selim et al., 2012). It is generally considered that phytotoxicity is eliminated when GI attains a level greater than 80% (Tiquia, 2010). In this maize seeds germination was inhibited by to a greater degree in the control set-up ie MSW control relative to the set ups. Generally germination index of values below 50% suggest that that phytotoxic compounds might have been fully metabolised thus

inhibiting germination (Selim et al., 2012). In case of MSW control the salt concentration might have inhibited the maize development.

Conclusion

It can be inferred from this study that the maturity parameters and germination index of the different compost indicates that though all the three animal manures could be used as suitable activators for the composting of municipal solid waste, poultry manure had the best activating properties amongst the three manures under the experimental conditions used in this study. These observations could be relevant for the improvement of municipal solid waste management in many developing economies. Furthermore the low technical and economical requirements of the modified Rapid Berkly method as used in this study could make it a promising option in unban localities.

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