

EVALUATION OF TOXICITY AND RESPONSE OF AFRICAN CATFISH (*Clarias gariepinus*) EXPOSED TO MANGANESE CHLORIDE SOLUTION

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ABSTRACT

This study investigated the acute toxicity and behavioural response of African catfish (*Clarias gariepinus*) to manganese chloride (MnCl₂) solution. One hundred (100) pieces of juvenile *Clarias gariepinus* were acclimatized for 14 days in ten (10) plastic bowls containing 25 L of water each. The acute-toxicity of manganese chloride, LC₅₀ and sub chronic exposure (28 days) were investigated. For sub-lethal exposure; fifty pieces of *C. gariepinus* were divided into five groups of ten in bowls labelled (A to E). The fish in groups (A to D) were exposed to manganese chloride solutions at: 68.5, 137, 205.5 and 274 mg/L respectively while group G served as the control. On day 28, the exposed fish were sacrificed; blood, muscle, liver, and gills were collected for haematological and histological studies. The activities of the superoxide dismutase (SOD) and glutathione-S-transferase (GST) were determined using standard methods. The manganese concentrations in the liver, gills and muscle tissues were determined using Atomic Absorption Spectrophotometer (AAS). The LC₅₀ obtained was 2.74 g/L. The mean water temperature showed that there was no significant variation at ($p < 0.05$) between the control and other groups. The results of growth parameters showed that fish exposed to 137 mg/L manganese concentration had better growth performance than those exposed to higher concentrations. The activities of SOD and GST in the liver of fish increased significantly ($p < 0.05$) as manganese concentration increases. The red blood cells concentrations decreases while white blood cells increases as the concentration of manganese increases respectively. The mean bioaccumulation of manganese in the organs of *C. gariepinus* followed the order of liver > gill > muscle. The histopathological study revealed alterations in the liver, gill and muscle tissues. This study concluded that manganese at concentration greater than 137 mg/L elicits adverse effect on *C. gariepinus* growth performance, haematology, histopathology as well as liver SOD and GST activities.

Key words: Behavioural response, Cat-fish, Juvenile, Manganese chloride, Toxicity

INTRODUCTION

Rivers and ponds are polluted due to the indiscriminate discharge of domestic, agricultural and industrial effluents which are toxic and contribute to the deterioration of water quality by changing its physicochemical nature leading to stress on aquatic organisms (Kamble and Tapale, 2011, Adesina *et al.*, 2018). Manganese (Mn) is the twelfth most abundant element which makes up about 0.1 % of the Earth's crust. It exists in a variety of oxidation states, Mn²⁺ and Mn³⁺ being the most biologically important.

Although the elemental (metal) form of manganese does not occur naturally in the environment, manganese is a component of over 100 minerals. Metal contamination of aquatic ecosystems has long been recognized as a serious

pollution problem (Baby *et al.*, 2010). The African catfish inhabits a wide range of water bodies including swamps, lakes and rivers; they thrive in harsh environmental conditions including muddy, turbid and oxygen depleted water bodies with the aid of their accessory air breathing organ (labyrinth organ) that allows them to breathe atmospheric oxygen. It has been observed that catfish from dirty ponds bioaccumulate metals, thus posing a great threat to the health of consumers (Farombi *et al.* 2007).

Manganese is an essential nutrient in animals; it plays a role in bone mineralization, protein and energy metabolism, metabolic regulation, cellular protection from damaging effect of free radical species and formation of glycosaminoglycans. In the last century, the massive production of

manganese-containing compounds (chemical products, alloys, steel, iron, ceramics and fungicide products) has attracted the attention of scientists who investigated manganese as a potential emerging contaminant in the environment, and especially in the marine environment (CICAD, 2004). In humans, excessive exposure to manganese has been shown to cause neurotoxicity, associated with a characteristic syndrome called 'manganese madness' or 'Parkinson-like' diseases (Perl and Olanow, 2007).

In humans, the neurological damage induced by excessive manganese exposure has been well documented for over a century (Takeda, 2003). Manganese toxicity in human could lead to serious neurological and reproductive disorders. The anthropogenic source of contamination is usually through industrial activities. It has been observed that *C. gariepinus* is one of the most cultured and consumed fish species in Nigeria and it bioaccumulates metals (including manganese) when cultured in polluted ponds. This study was designed to evaluate the toxicity of African catfish *C. gariepinus* exposed to manganese chloride solution.

MATERIALS AND METHODS

Fish Collection and Acclimatization

One hundred (100) 8-10 weeks old *C. gariepinus*; 30-45g belonging to a single population were purchased from the hatchery unit, Department of Fisheries and Aquaculture, Obafemi Awolowo University, Ile-Ife, Nigeria. The test organisms were kept in a large plastic container and acclimatized for a period of fourteen days. During acclimatization, renewal bioassay was employed and fish were fed twice daily with fish feed (Copens) of about 40% crude protein content. The water used was obtained from hand dug well in the Faculty of Agriculture, Obafemi Awolowo University, Campus, Ile-Ife.

Reagents and Chemicals

The reagents used were of analytical grade. They were obtained from various sources such as Sigma Fine Chemicals Limited, Sweden, British Drug House Limited, London and Oxford Laboratory Reagents, Mumbai, India

TOXICOLOGICAL INVESTIGATION

Toxicity Range Finding Test

The range finding test was carried out to determine the definitive concentrations of manganese to be used for the toxicity evaluation according to the procedure described by ASTM (2007). Four (4) test solutions were prepared in stock bowls labelled A-D containing varying concentrations of manganese chloride and a control. Ten (10) fish were kept in each bowl and observed for mortality after 24, 48, 72, and 96 hr as described by Reish and Oshida, (1987). Toxicity range of manganese concentration to *C. gariepinus* was then estimated from the mortality results.

Acute Toxicity Study and LC₅₀ Determination

The results obtained from the toxicity range finding test was used as a guide for the acute toxicity test. Four (4) different concentrations of manganese chloride were prepared in four plastic bowls (2.0, 3.0, 4.0 and 5.0 g/L) and control, this set was used for the 96 hr acute toxicity studies in a static exposure system (Obuotor, 2004). Five (5) *Clarias gariepinus* juveniles (50 ± 4 g) were introduced carefully into each of the bowls containing the test solutions in duplicate. The behavioural response of fish to varying concentrations of manganese chloride was observed as well as mortality in each of the exposure chambers and these were recorded after 12, 24, 48, 72, and 96 hr. The 96 hr LC₅₀ value was then calculated from the data obtained using probit method (Finney, 1971).

Sub chronic Toxicity Study

Fifty (50) *C. gariepinus* were divided into five groups (A to E) in duplicates and the experimental fish was exposed to test solutions containing the following manganese concentrations: 68.5, 137, 205.5 and 274 mg/L which represent 2.5, 5.0, 7.5 and 10% of LC₅₀ respectively), group E is the control. Static/renewal bioassay was employed in the experiment and the test solutions were renewed every 48 hr with freshly prepared solutions in order to maintain stable concentration of the metal and to minimize waste accumulation.

Physicochemical parameters of the water used were determined and recorded; they are dissolved

oxygen, pH, temperature and conductivity. Careful observations were made to determine the growth response and toxicity of the solution to *C. gariepinus*. The fish were fed at the rate of 5% of their body weight between the hours of 7.00 hr - 8.00 hr and 18.00 hr-19.00 hr during the exposure period that lasted for four weeks (28 days). The feed was withdrawn 24 hrs to the termination of the experiment in accordance with Achionye-Nzeh *et al.* (2004) method. Data obtained were analysed with appropriate statistical methods.

Liver homogenate was prepared as described by Babalola and Areola, (2010).

Assay of Glutathione S-Transferase (GST) and Superoxide Dismutase (SOD) in the Liver Homogenate of *C. gariepinus*

Glutathione S-transferase activity was determined by following the method of Habig *et al.* (1974) while the SOD assay was carried out as described by Fridovic and Maccord (1969).

Haematological Analysis

Fish were removed from each tank and cleaned with a wet towel. Blood was then collected by making an incision on the caudal peduncle of the fish; the flow of blood was directed into well labelled bottles of potassium ethylene-diamine-tetraacetic acid (K+EDTA). The blood samples were analysed at the Department of Haematology, College of Health Science, Obafemi Awolowo University, Ile-Ife.

Tissue Digestion for AAS Analysis

One (1 g) each of the separate tissues (liver, gills and muscle) was placed in a test tube; 5 ml of concentrated nitric acid was carefully added to each tube, covered with cotton wool and left on the bench overnight. The following day, the wool was removed and the reacting mixture in tubes were boiled in water bath for 1 hr. The tubes were

removed from water bath and allowed to cool and then diluted to 25 ml with deionised water. The solution obtained was used for manganese analysis. The concentrations of manganese were determined using PG 990 Atomic Absorption Spectrophotometer.

Histopathology Study

The histopathology study was done in the Department of Anatomy and Cell Biology, Obafemi Awolowo University, Ile-Ife.

Statistical Analysis

The statistical package used for data analysis was IBM SPSS Inc version 20.0 and Microsoft Excel 2013 version. Statistical data obtained from this study were presented as mean \pm SEM (Standard Error of Mean) and the significant mean differences were separated at 0.05 probability levels.

RESULTS

Physico-chemical Quality of Culture Media

The mean values of physico-chemical parameters of the culture media are summarized in table 1. The highest weekly temperature value recorded was 26.02 ± 1.11 °C while the lowest weekly water temperature recorded was 25.82 ± 0.89 °C. The variation in water temperature showed that the groups containing manganese were not significantly different ($p < 0.05$) from the control. Changes in pH of the culture water showed that the group with the highest concentration of manganese had the highest mean pH of 7.54 ± 0.62 while the control had the lowest mean pH of 6.69 ± 0.31 . It was observed that as the concentrations of manganese increases, the water pH also increases. The changes in pH showed that the water pH of the groups treated with manganese were not significantly different ($p < 0.05$) from the pH of the control.

Table 1: The mean values of Physico-chemical Parameters of the Culture media

Parameters	Concentrations of Manganese Chloride (mg/L)				
	Control	Group A	Group B	Group C	Group D
Temperature (°C)	26.05± 1.21	25.82 ± 0.89	25.85± 1.25	26.02± 1.11	26.01± 1.26
pH	6.69± 0.31	7.0± 0.21	7.13± 0.27	7.36 ± 0.34	7.54 ± 0.62
DO (mg/L)	6.93 ± 0.22	6.95 ± 0.53	6.72 ± 0.22	6.70 ± 0.36	6.47 ± 0.28
Conductivity (µs/cm)	5.31 ± 0.11	11.14 ± 0.25	13.92 ± 0.59	16.29 ± 1.55	20.67 ± 1.06

Values are expressed as Mean ± SEM, n=5

Concentrations of manganese chloride: A= 68.5 mg/L, B = 137, C = 205.5 mg/L, D = 274 mg/L, DO- Dissolved oxygen

The mean concentration of dissolved oxygen (DO) of the culture water obtained for all the groups ranges from 6.95 ± 0.53 mg/L to 6.47 ± 0.28 mg/L. The results obtained showed that there was no significant difference ($P > 0.05$) in dissolved oxygen of water of fish exposed to manganese however a significant decrease ($P < 0.05$) was observed between dissolved oxygen of water in the control and the group with highest manganese concentration.

The weekly conductivity of the culture water ranged from 4.59 to 23.52 µs/cm. The group with the highest manganese concentration had the highest conductivity value. Generally, it was observed that as the manganese concentration increases, the mean conductivity values also increases.

Behavioural Changes Observed in *C. gariepinus* During Acute Toxicity Study

Behavioural changes observed during acute toxicity test include loss of reflex, moulting, discolouration, air gulping, erratic swimming, barbell deformation and excessive mucus secretion in the test organisms. These parameters were monitored and recorded for the varying concentration of manganese. Loss of reflex was observed only in group D with highest concentration of Mn within 24 hr of exposure. However, at 96 hr of acute toxicity test, group B, C and D showed loss of reflex. The behavioural responses observed increased with the metal concentration as shown in table 2. The table also confirmed other behavioural changes noticed in the experimental organisms exposed to varying concentration of manganese.

Table 2: Behavioural Changes Observed in *C. gariepinus* During Acute Toxicity Study

Exposure Time (hr)	24					48					72					96				
	0	2	3	4	5	0	2	3	4	5	0	2	3	4	5	0	2	3	4	5
MnCl ₂ (g/l)																				
Behaviour																				
Loss of reflex	-	-	-	-	+	-	-	-	+	+	-	-	-	+	+	-	-	+	+	+
Moulting	-	-	-	+	+	-	-	+	+	+	-	-	+	+	+	-	+	+	+	+
Discolouration	-	-	-	-	-	-	-	-	-	+	-	-	-	+	+	-	-	+	+	+
Air Gulping	-	-	-	-	+	-	-	-	+	+	-	-	+	+	+	-	+	+	+	+
Erratic Swimming	-	-	-	+	+	-	-	+	+	+	-	-	+	+	+	-	+	+	+	+
Barbel Deformation	-	-	-	-	-	-	-	-	+	+	-	-	+	+	+	-	-	+	+	+
Excessive Mucous Secretion	-	-	-	-	-	-	-	-	-	-	-	-	-	+	+	-	-	-	+	+

Key: + = Present, - = Absent

Mortality during Acute Toxicity Study

During the 96 hr acute toxicity study, mortality was recorded in all the treatment groups. Fish in Group A, B, and C had no mortality at 24 hr of exposure while group D with the highest concentration of manganese had 20% mortality. At 48 hr of exposure, mortality rate rose to 60% in group D, 20% in group C and no mortality was

observed in group B and A. Group D had mortality rate of 80% at 72 hr of acute toxicity test while mortality rose to 40% in group C, 20% for group B and no record of mortality in Group A. Nevertheless, mortality rose to 100% in group D, 80% in group C, 60% in group B and 20% in group A as shown in table 3.

Table 3: Mortality of Fish during Acute Toxicity Study.

Time of Exposure(hrs)	Control	A	B	C	D
24	-	-	-	-	1
48	-	-	-	1	2
72	-	-	1	1	1
96	-	1	2	2	1
Total Mortality	0/5	1/5	3/5	4/5	5/5
Total Mortality (%)	0	20	60	80	100

Manganese chloride concentration: Group A (2 g/L), Group B (3 g/L), Group C (4 g/L), Group D (5 g/L).

Determination of LC₅₀

The lethal concentration (LC₅₀) was calculated from the graph of manganese concentrations against mortality. The LC₅₀ value obtained from the graph was 2.74 g/L.

Sub-chronic Toxicity Study

The fish mortality was expressed as the number of dead fish per bowl in the control and the groups with manganese throughout the exposure period. There was no mortality during the 28 days sub-chronic test in treatment bowls and the control, survival rate was 100%.

Growth Performance Indices

The growth performance of *C. gariepinus* cultured in varying concentration of, manganese under renewal biostatic system for 28 days is presented in table 4. There was no significant difference ($P > 0.05$) in the initial weight (Wi) of fish exposed to manganese and the control. The mean weight gained (MWG) decreases from group A with

MWG of 34.00 ± 0.53 g to group D with MWG of 22.10 ± 0.72 g. As the manganese concentration increases, the MWG decreases. There was a significant difference ($P < 0.05$) in mean weight gained between the fish in the control group and the fish cultured in varying concentration of manganese. Specific growth rate (SRG) values decrease as the concentrations of manganese increases, fish in control group had the highest SGR while the fish with highest concentration of manganese had the lowest SGR. Feed conversion ratio (FCR) was found to increase as the concentration of manganese increases; as a result of this, the control group had the lowest FCR while the group with highest manganese concentration had the highest FCR (table 4). Comparison of the FCR values between the control group and groups with manganese showed that the difference between them were statistically significant ($P < 0.05$). FCR is therefore inversely proportional to SGR, as FCR increases; SGR decreases (figure 1).

Table 4: The Growth and Performance of *C. gariepinus* in Different Concentration of Manganese

Index	Control	A	B	C	D
W _i (g)	50.20 ^a ±0.50	49.80 ^a ±0.45	49.50 ^a ±0.75	52.50 ^a ±0.55	51.50 ^a ±0.85
W _f (g)	86.60 ^a ±0.48	83.80 ^{ab} ±0.69	80.80 ^b ±0.46	80.60 ^b ±0.59	73.60 ^{bc} ±0.74
MWG (g)	36.40 ^a ±0.58	34.00 ^b ±0.53	31.40 ^{ab} ±0.39	28.10 ^b ±0.48	22.10 ^c ±0.72
SGR	2.51 ^a ±0.23	1.86 ^{ab} ±0.28	1.71 ^{ab} ±0.43	1.68 ^{ab} ±0.33	1.40 ^c ±0.27
FCR	1.93 ^b ±0.37	2.05 ^b ±0.56	2.21 ^b ±0.46	2.61 ^{ab} ±0.43	3.26 ^a ±0.09
SR (%)	100	100	100	100	100

Values are expressed as Mean ± SEM, n=5

Manganese chloride concentrations: A= 68.5 mg/L, B = 137, C = 205.5 mg/L, D = 274 mg/L. Abbreviations: W_i = Initial body weight, W_f = Final body weight, MWG = Mean Weight gained, FCR= Feed Conversion Ratio, SGR = Specific Growth Rate and SR = Survival Rate.

Values with subscripts are statistically different at p < 0.05 when compared with the control group.

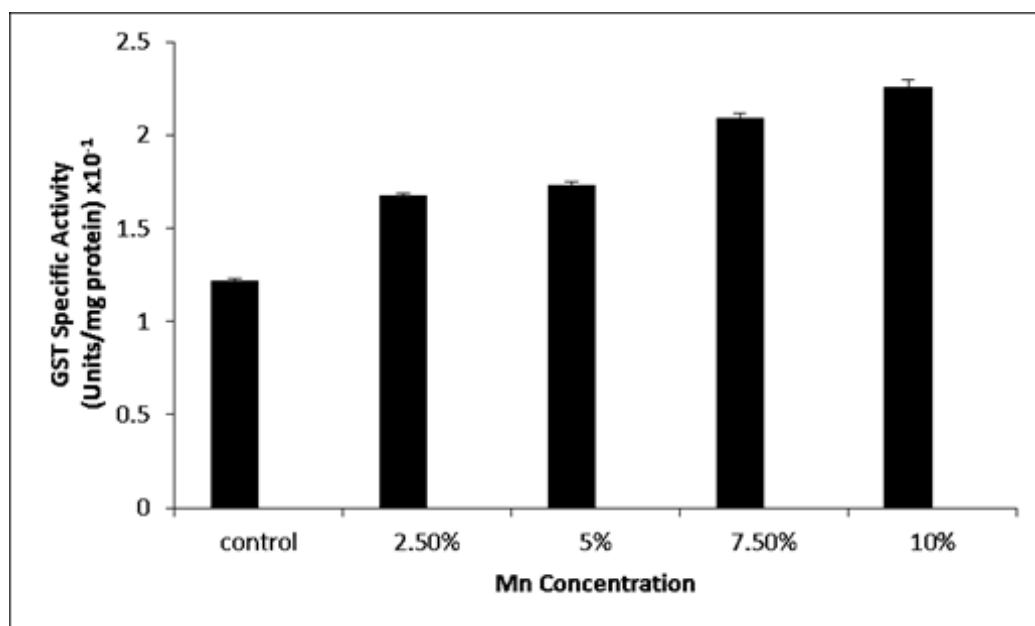


Figure 1: Activity of GST in liver of *C. gariepinus* on exposure to various concentrations of manganese for a period of 28 days. The results are represented as Mean ± SEM, n=5.

[MnCl₂] 2.5% = 68.5 mg/L, 5% = 137 mg/L, 7.5% = 205.5 mg/L and 10% = 274 mg/L

Effects Manganese Exposure on Haematological Parameters of *C. gariepinus*

The results of haematological parameters of *C. gariepinus* subjected to sub-chronic concentrations of manganese and the control are shown in table 5. The results showed that white blood cell count in the experimental fish increase as manganese concentration increases. The white blood counts were not significantly different from the control at (P<0.05). However, at a manganese concentration greater than 137 mg/L, it showed a significant variation at (P< 0.05) from the control.. Decrease in the red blood cell (RBC) values were also observed as the concentration of manganese increases. Although, the difference between the

control and fish cultured in manganese were not significantly different in RBC values except at 274 mg/dL (P<0.05).

Furthermore, the fish exposed to 68.5 mg/L had haemoglobin (HGB) concentration that was not significantly different (P > 0.05) from the control. However, the fish in other manganese concentrations showed a decrease in the haemoglobin concentration and the decreases were significantly different (P < 0.05) from the control group. The values obtained for mean concentration of haemoglobin (MCH) and platelet (PLT) followed the same trend in all the groups as shown in table 5.

Table 5: Effect of Manganese on Haematological Parameters of *C. gariepinus* after 28 Days Exposure

Parameters	Manganese Treatment (Dose)				
	Control	A	B	C	D
WBC($10^3/\mu\text{L}$)	50.11 ^a ±0.14	52.32 ^a ±0.08	51.59 ^{ab} ±0.23	58.91 ^b ±0.17	62.83 ^b ±0.19
RBC($10^6/\mu\text{L}$)	2.73 ^a ±0.22	2.48 ^{ab} ±0.32	2.43 ^{ab} ±0.29	2.41 ^{ab} ±0.11	2.21 ^b ±0.16
HGB(g/dL)	12.33 ^a ±0.45	12.27 ^a ±0.32	11.42 ^{ab} ±0.24	10.76 ^b ±0.29	9.29 ^b ±0.42
LYM($10^3/\mu\text{L}$)	94.40 ^a ±0.23	94.90 ^a ±0.15	96.5 ^b ±0.09	96.4 ^b ±0.37	94.2 ^a ±0.38
MCH(pg)	46.93 ^a ±0.09	45.10 ^{ab} ±0.43	46.90 ^a ±0.22	44.30 ^{ab} ±0.38	41.60 ^b ±0.14
MCHC(pg/dL)	38.66 ^a ±0.25	34.41 ^{ab} ±0.14	32.18 ^{ab} ±0.19	31.64 ^a ±0.21	31.74 ^a ±0.12
PLT($10^3/\mu\text{L}$)	57.45 ^a ±0.28	42.39 ^a ±0.22	46.43 ^{ab} ±0.29	48.35 ^{ab} ±0.31	41.57 ^a ±0.21

Values with different subscripts are significantly different at $p < 0.05$. Values are expressed as Mean \pm SEM, $n=5$,

Manganese chloride concentration: A = 68.5mg/L, B = 137mg/L, C = 205.5mg/L and D=274mg/L

Abbreviations: RBC = Red Blood Cell Count, WBC = White Blood Cell Count, PLT = Platelet Count, HBG = Haemoglobin, MCH = Mean Corpuscular Haemoglobin, MCHC = Mean Corpuscular Haemoglobin Concentration.

Glutathione transferase Activity (GST)

The values of GST activity revealed that there was a significant variation ($p < 0.05$) in GST specific activity in liver of the juvenile *C. gariepinus* between the control and the treatment groups. The values revealed that the specific activity of GST in the liver of fish exposed to various manganese concentrations were significantly higher ($p < 0.05$) than GST activity in the liver of the fish in the control as shown in figure 1.

Superoxide Dismutase (SOD) Activity

The SOD activity increased as the concentration of manganese increases in the *C. gariepinus* liver. The difference in SOD activity at lower concentrations of manganese were not significant but at higher concentrations (205.5 and 274 mg/L), there were significant differences at $p < 0.05$ when compared with the control (Figure 2).

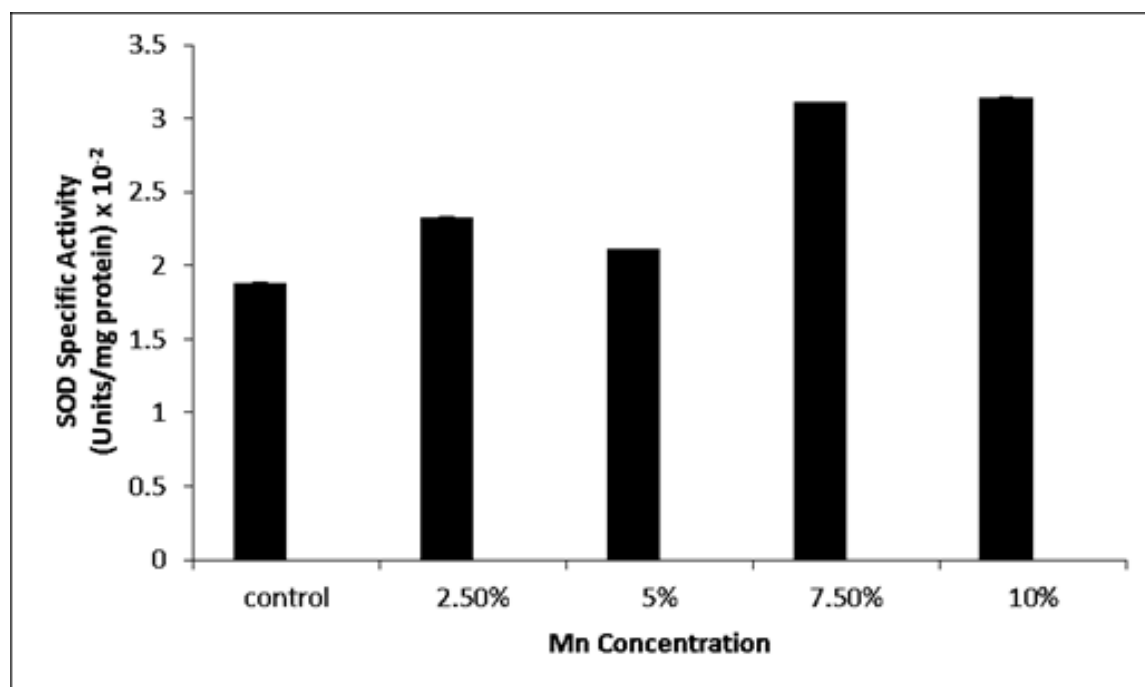


Figure 2: Activity of SOD (U/mg protein) in Liver *C. gariepinus* Exposed to Various Concentrations of Manganese for a Period of 28 days.

The results are represented as Mean \pm SD, $n=5$.

[MnCl₂] 2.5% = 68.5 mg/L, 5% = 137 mg/L, 7.5% = 205.5 mg/L and 10% = 274 mg/L

Analysis of Manganese Concentrations in Tissues of *C. gariepinus*

The graphical representation of concentrations ($\mu\text{g/g}$) of manganese in organs of *C. gariepinus* exposed to sub-chronic concentrations of manganese for 28 days is shown in figure 3. The liver of fish exposed to 274 mg/L manganese had the highest bioaccumulation of $2.34 \pm 0.039 \mu\text{g/g}$ of manganese followed by the gill ($1.65 \pm 0.108 \mu\text{g/g}$) while the muscle had the lowest concentration of $1.39 \pm 0.163 \mu\text{g/g}$. In groups with lower concentration of manganese (68.5

mg/L, 137 mg/L), the rate of manganese bioaccumulation was found to be in the order of gill > muscle > liver while fish exposed to higher concentration of manganese (205.5 mg/L, 274 mg/L) was in the order of liver > gill > muscle. The difference in the bioaccumulation in the various organs/tissues however did not differ significantly at ($P < 0.05$) in fish exposed to lower concentration of Mn but there was a significant difference in the rate of bioaccumulation of manganese in organs of fish exposed to 205.5 mg/L and 274 mg/L.

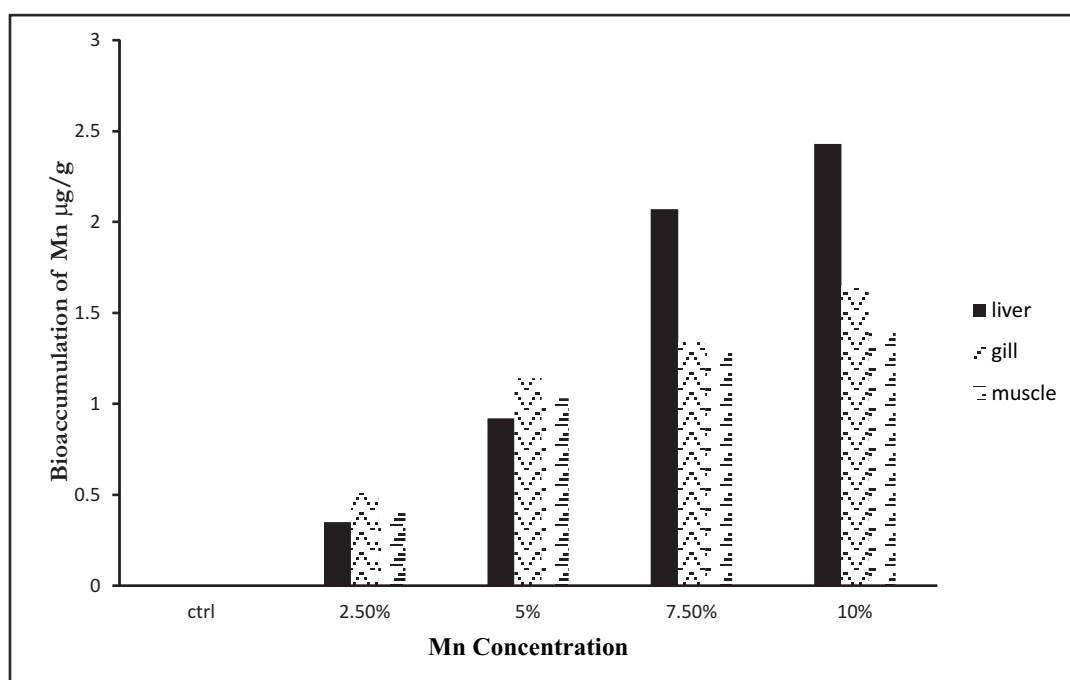


Figure 3: Bioaccumulation of Manganese in Organs of *C. gariepinus*

Histopathology

The histopathology of different *C. gariepinus* tissues revealed that there were several histopathological changes in different organs/tissues (liver, gills, and muscle) of the fish subjected to varying concentrations of manganese as shown in plates 1 and 2

Histopathological Changes in the Liver

In the present investigation, the photomicrograph of liver of the control fish exhibited a normal architecture with hepatocytes presenting a homogenous cytoplasm and a spherical nucleus (N) (Plate 1). The liver of fish exposed to 68.5 mg/L of manganese exhibited focal area of necrosis (NC), while the liver of fish exposed to 137 mg/L and 205.5 mg/L of manganese showed

liver anomalies such as degenerative nuclei (DN), cytoplasmic vacuolation (CV) and necrosis (NC) (Plates 1b and 1c). In plate 1d, the photomicrograph of liver of the fish exposed to 274 mg/L of manganese was seen to have severe abnormalities such as irregular shaped hepatocytes and nucleus in a lateral position. Also, hepatocytes were located among blood capillaries called sinusoids (SS) forming cord-like structures which indicate vascular dilation (D), congestion and dilation of sinusoids (CS and DS).

Histopathological Changes in the Gills

The gill is made up of filaments of primary lamellae arranged in double rows. Secondary lamellae arise from these filaments. The secondary

lamellae are lined by a squamous epithelium. Plate 2 shows the control gill tissues which consisted of primary lamellae (PL) and secondary lamellae (SL). The secondary lamellae composed of a single layer of epithelial cells. There were no conspicuous changes in the gill tissues of the control fish. However, plates 2a and b shows the gill tissues of *C. gariepinus* exposed to 68.5 mg/L and 137 mg/L of manganese. Shortening of secondary lamellar was observed in plate 2a, while

plate 2b showed fusion of secondary lamellae, degeneration of the epithelia and mild epithelial vacuolation of the secondary lamellae. Plate 2c and d represents the photomicrograph of gill tissues of fish exposed to 205.5 mg/L and 274 mg/L of manganese respectively, plate 2c reveals degenerating epithelia and hyperplasia of the secondary lamellae while plate 2d shows aneurysm around the primary lamellae and epithelia of the secondary lamellae

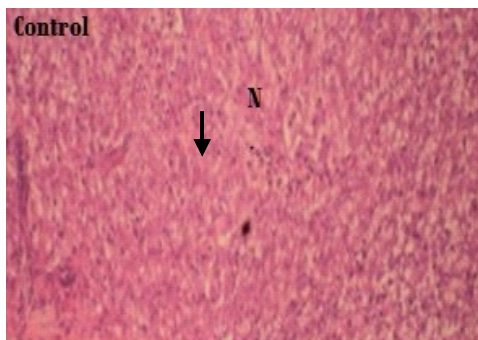


Plate 1: Control

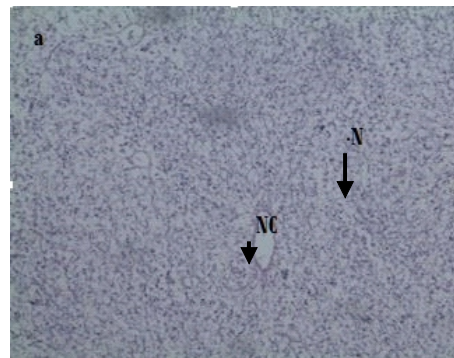


Plate 1a

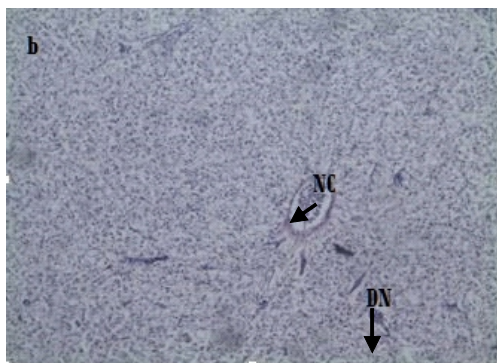


Plate 1b

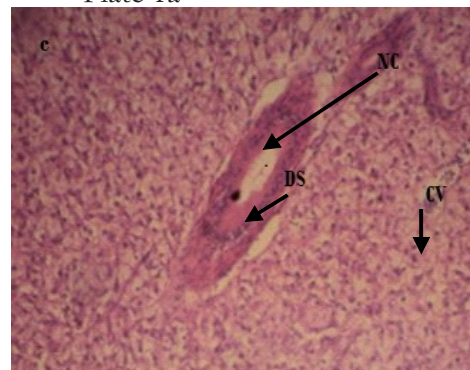


Plate 1c

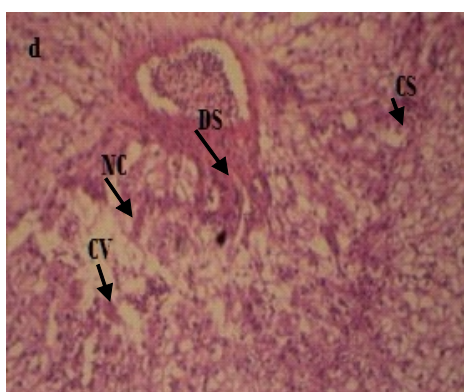


Plate 1.d

Photomicrograph of Liver Section of Rat (Plates a, b, c and d) Exposed to 68.5, 137, 205 and 274 mg/L of $MnCl_2$ respectively, together with the Control (mag. x 100).

Abbreviations: Spherical Nucleus (N), Area of Necrosis (NC), Degenerative Nuclei (DN), Cytoplasmic Vacuolation (CV), Vascular Dilatation (D) and Congestion and Dilatation of sinusoids (CS and DS).

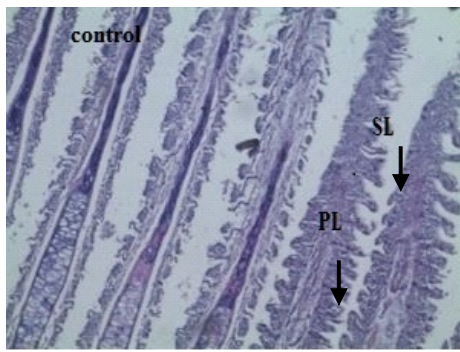


Plate 2: Control

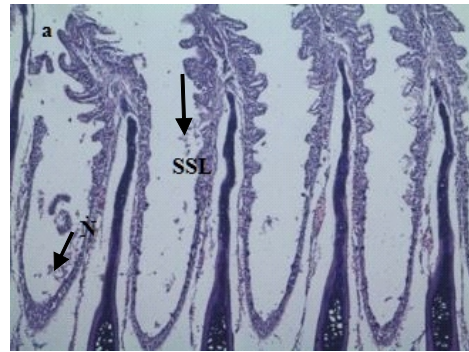


Plate: 2a

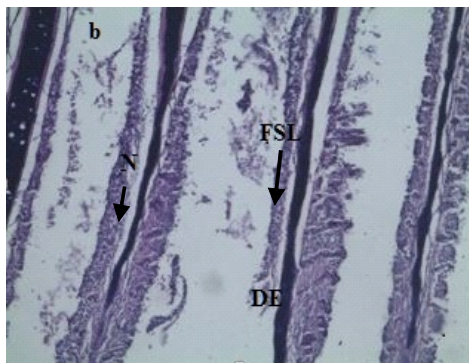


Plate: 2b

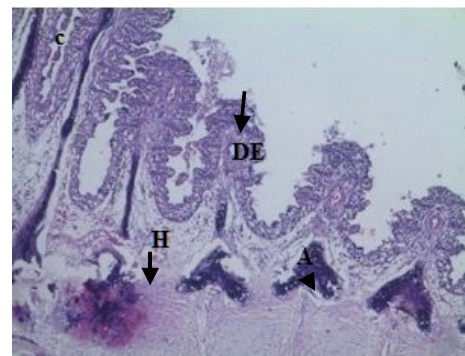


Plate: 2c

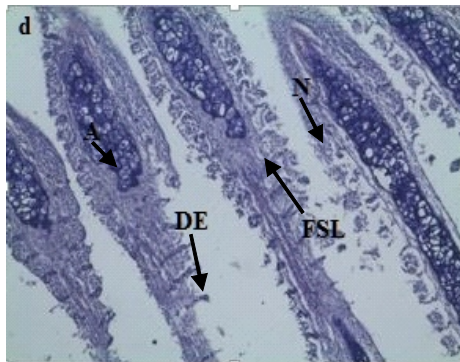


Plate: 2d

Photomicrograph of Gill Section of rat (Plates a, b, c and d) Exposed to 68.5, 137, 205 and 274 mg/L of $MnCl_2$ respectively together with the Control (mag, x100).

Abbreviations: Primary Lamellae (PL), Secondary Lamellae (SL), Fusion of Secondary Lamellae (FSL), Shortening of Secondary Lamellae (SSL), Hyperplasia (H), Degenerative Epithelium (DE), Focal Area of Necrosis (N), Fusion of Secondary Lamellae (FSL).

DISCUSSION

Water quality plays a major role in the growth and gross fish production. Owing to its significance, fish farmers always keep these parameters under consideration when siting their fish ponds. The results of physicochemical analysis of the culture water containing varying concentrations of manganese showed that pH increases with increase in manganese concentrations.

Nevertheless, the pH of the water decreases with time of exposure (day 1 to 3) in all the groups exposed to manganese. Temperature has a profound effect on biological processes; the metabolic activity of aquatic organisms also increases with increase in temperature (Howerton, 2001). The normal range of temperature in the tropics to which fish is adapted is 22-35 °C (WHO, 2008). The mean temperature recorded during this

study was within the recommended limit for fresh water fishes. Also, there was no significant difference ($p < 0.05$) between the control and the groups exposed to manganese.

Studies have shown that suitable water quality for any fish culture in the tropical region must have dissolved oxygen concentration of at least 3 mg/L (Robert, 2007). The mean dissolved oxygen contents recorded in this study was (6.47 – 6.95 mg/L) greater than 5 mg/L in all the treated groups. In this study, the dissolved oxygen in water of the manganese groups were significantly different ($p < 0.05$) from the control group. This result however contradicts the report of Khalid, (2011) who observed a decrease in dissolved oxygen concentration with increase in the level of nickel in culture water. The regular and constant changing of culture water may be partly responsible for the high DO recorded in this study. This observation revealed that DO of culture water can be maintained through regular changing of the culture water.

The 96 hr LC_{50} value obtained for *Clarias gariepinus* exposed to varying concentrations of manganese in this study was 2.74 g/L, this value falls within the range of manganese concentration that have been documented in a number of fish species by different authors. For example, Garg *et al.* (1989) reported LC_{50} of 3.35 g/L in Indian catfish (*Heteropneustes fossilis*) and 3.01 g/L in Indian freshwater murrel (*Channa punctatus*). Alteration in behaviour is considered as a sensitive biomarker to evaluate the toxicant exposure and effect (Gerhardt, 2007). The studies on fish behaviours provide a lot of knowledge and information because any behaviour alteration can be related to physiological biomarker in aquatic species. The behavioural changes recorded during the acute exposure of *C. gariepinus* to varying concentration of manganese include: loss of reflex, moulting, discolouration, air gulping, erratic swimming, barbell deformation and excessive mucus secretion. However, deep behavioural changes observed in fish exposed to higher concentration of manganese showed that it could induce oxidative stress (Javed, 2012). There are many ways that behaviour and physiology can interact to affect aquatic toxicology. A few studies have shown correlations between behavioural and

physiological indicators of toxicity and have therefore succeeded in eliminating the complicating effects faced when comparing different behavioural and physiological studies. Affected fish with behavioural alteration toward toxicant especially pesticides and heavy metals has also been reported (Javed, 2012).

A significant decrease in growth parameters were recorded in fish exposed to manganese concentration when compared with the control. Stress-induced reduction in growth is a well-recognized toxic effect in fish; such effects on growth are expected when organisms are exposed to chemical pollutants, if the exposure is sufficient to produce a stress response. This is the importance of evaluating growth response and oxidative stress in commercially important fish species (Adeogun *et al.*, 2011). The significant variation ($p < 0.05$) in the values of specific growth rate (SGR) of the fish could be attributed to the toxic effects of manganese. The observed increase in FCR of fish exposed to manganese indicated the inability of the fish to convert feed consumed into body protein. This study has revealed that fish exposed to manganese concentration greater than 137 mg/L will experience poor growth.

The haematological observation showed that the fish cultured in lower manganese concentration had better blood health status than those of higher concentrations. However all the haematological parameters measured in this study were within the recommended physiological range reported for *C. gariepinus*. The blood parameters- RBC and HGB count decreased with increasing concentration of manganese and this become significantly lower ($P < 0.05$) at higher concentration when compared with the control. The variation observed in the haematological parameters of fish in this study could be as a result of the toxic effect of manganese. In addition, Roels *et al.* (1992) have reported adverse haematological effects following occupational exposure to manganese. Oyawoye and Ogunkunle, (1998) reported an increase in white blood cells and lymphocyte count which is usually associated with microbial infection or the presence of foreign bodies. The haematological results obtained from this study revealed that exposure of catfish (*Clarias gariepinus*) to manganese concentration greater than 205.5

mg/L would have adverse effect on the fish haematological parameters.

The antioxidant enzymes activities were found to increase as the concentration of manganese increases in all the groups. This observation could be due to the fact that exposure of fish to pollutants could trigger the production of antioxidants to overcome stressful conditions generated by such pollutants. Super oxide dismutase (SOD), provides an important means of cellular defence against free radical damage, therefore this could be responsible for higher activities of SOD observed in the liver of fish in groups exposed to manganese. However, this finding conflicts with a report by Olagoke, (2008) who reported a decrease in activities of SOD in fish exposed to polycyclic aromatic hydrocarbons stating that an inhibition of the enzyme SOD will expectedly result in a reduction in the activity of catalase due to a decrease in H₂O₂ generation from SOD activities. The GST activities in the liver of the fish exposed to higher manganese concentrations had higher activities when compared with GST activities of fish in the control and those of lower concentrations (2.5 and 5%). These results are in accordance with the works of Mannervik and Guthenberg (1981) on white fish tissues.

The analysis of metal concentrations in the liver, gill and muscle, showed a significant ($p < 0.05$) difference in the heavy metal concentrations across the organs/tissues of *C. gariepinus*. The liver of the fish had the highest mean concentrations followed by the gills, while the muscle had the lowest metal concentration. The concentrations of manganese in tissues are different because the rate of metal bioaccumulation is not the same (Karadede *et al.*, 2004). In this study, the metal levels in fish were found to be lowest in the muscle when compared with liver and gills.

The histopathological examination revealed that, the liver of fish exposed to 5, 7.5 and 10% concentration of manganese revealed degeneration of the hepatocytes, congestion of central vein, necrosis (NC), degenerative nuclei (DN), cytoplasmic vacuolation (CV), vascular dilation (D) and sinusoid congestion (CD) and dilation of sinusoids (DS) in the hepatic cells when

compared with the normal hepatic cell in control fish. These results were in accordance with those reported by Van Dyk (2003).

Gill is the first direct contact with water from external environment and changes in fish gill is the most usually distinguished reactions to environmental toxins. In a study by Olojo *et al.* (2004), it was reported that necrosis and desquamation of gill epithelium as well as lamellar curling and aneurisms were the direct deleterious effects reported in chronic lead exposed *C. gariepinus*. The gill of fish in the control are normal while the gills of fish exposed to manganese showed structural deformation such as epithelial lifting at secondary lamella, hyperplasia of primary epithelium, fusion of secondary lamella, aneurisms, necrosis and infiltration of inflammatory cells with the disintegrate of epithelial cells of secondary lamellae including mucus secretion. It was also reported that edema of the gill epithelium is one of the main structural changes caused by exposure to heavy metals (Mallatt, 1985).

CONCLUSION

This study concluded that manganese concentration greater than 137 mg/L in the pond water had adverse effects on the growth performance generally, therefore such water is not suitable for *C. gariepinus* farming.

RECOMMENDATION

Discharge of industrial waste water containing manganese should be regulated to prevent its accumulation in aquatic organisms. Fish farmers should always check the manganese concentrations of their ponds before transferring fish into them. It is also necessary to carry out periodic checks on the quality of fish caught in polluted ponds to prevent human consumption of fish which had been contaminated with metals.

REFERENCES

- Achionye-Nzeh, C.G., Ogidiolu, O. and Salami, S. 2004. Growth response of *Clarias anguillaris* fingerlings fed diets formulated with *Macrotermes nigerensis*. *Journal of Pure and Applied Sciences*. 19: 1570-1573.
- Adeogun, A. O., Chukwuka, A. V. and Ibor, O.R. 2011. Impact of abattoir and saw-mill

- effluents on water quality of upper Ogun river (Abeokuta). *American Journal of Environmental Science*, 7(6): 525-530.
- Adesina, I.M., Bisi-Johnson, M., Aladesanmi, O.T., Okoh, A.I. and Ogunfowokan, A.O. 2018. Concentrations and human health risk of heavy metals in rivers in southwest Nigeria. *Journal of Health and Pollution*, 18(19): 1-14.
- ASTM, 2007. Standard guide for conducting acute toxicity tests on test materials with fishes, *Macroinvertebrates and Amphibians*, 11: 6-11.
- Babalola, O. O. and Areola, J. O. 2010. Interactive roles of terpenoid extract from the leaves of neem plant (*Azadirachta indica*, A. juss) on lead induced toxicity in pregnant rabbits. *Journal of Medicinal Plants Research*, 4(12): 1102-1107.
- Baby J., Raj, J.S., Biby, T.E., Sankargarnesh, P. And Rajan, S.S. 2010. Toxic effects of heavy metals on aquatic environment. *International Journal of Biological and chemical Science*, 4(4): 939-952.
- CICAD, 2004. Manganese and its compounds: Environmental aspects. Concise international chemical assessment. http://www.who.int/ipcs/publications/cicad/cicad63_rev_1.pdf
- Farombi, E.O. Adelowo, O.A and Y.R. Ajimoo, Y.R., 2007. Biomarker of oxidative stress and heavy metals as indicators of pollution in African catfish (*C. gariepinus*) from Ogun river, Nigeria. *International Journal of Environmental Research and Public Health*, 4 (2): 158-165.
- Finney, D. J. 1971. Probit Analysis. 3rd ed. Cambridge University press, Cambridge, England. Pp. 333.
- Fridovich, J and MacCord, J.M. (1969). Superoxide Dismutases., an enzyme function for erythrocyte hemocuprein (hemocuprein) *Journal of Biological Chemistry*, 24 (22): 6049-6055.
- Garg V. K., Garg, S. K. and Tyagi, S. K. 1989. Manganese induced hematological and biological anomalies in *Heteropneustes fossilis*. *Journal of Environmental Biology*, 10: 249-35.
- Gerhardt, A. 2007. Aquatic behavioural ecotoxicology - Prospects and limitations. *Human and Ecological Risk Assessment*, 13: 481-491.
- Habig, W. H., Pabst, M. J. and Jakoby, W. B. 1974. Glutathione-S-transferase. The first enzymatic step in mercapturic acid formation. *Journal of Biological Chemistry*, 49: 7130-7139.
- Howerton, R. 2001. Best management practices for Hawaiian aquaculture. Center for tropical and subtropical aquaculture, University of Hawaii, Sea Grant extension services, pp. 148.
- Javed, M. 2012. Growth responses of fish under chronic exposure of waterborne and dietary metals. *International Journal of Agricultural Biology*, 14: 281-285.
- Kamble, S. M. and Tapale, B. K. 2011. Effects of sublethal concentrations of a household detergent on certain biochemical constituents of catfish, *Mystus seenghala*. *Bioscience and Biotechnology Resources Communication*, 4: 198-204.
- Karadede, H., Oymak, S.A. and Ünlü, E. 2004. Heavy metals in mullet, liza Abu and cat fish, *Silurus triostegus*, from the Atatürk Dam Lake (Euphrates), Turkey. *Environment*, 30: 183-188.
- Khalid, A. G. 2011. Impact of nickel on haematological parameters and behavioural changes in *Cyprinus carpio* (common carp). *African Journal of Biotechnology*, 10(63): 13860-13866.
- Mallatt, J. 1985. Fish gill structural changes induced by toxicants and other irritants: A statistical review. *Canadian Journal of Aquatic Science*, 42: 630-648.
- Mannervik B, and Guthenberg C. 1981. Glutathione-S-Transferase (human placenta). *Methods in Enzymology*, 77: 231-235.
- Obuotor, E. M. 2004. The mode of action of Ichthyotoxic principles in *Raphia bookeri* fruit. Ph. D. Thesis, Obafemi Awolowo University, Ile-Ife. pp. 172.
- Olagoke, O. 2008. Lipid peroxidation and antioxidant defense enzymes in *Clarias gariepinus* as useful biomarkers for monitoring exposure to polycyclic aromatic hydro-carbons. M.Sc Theses, University of Lagos, Lagos, Nigeria. pp70.
- Olojo, E. A., Olurin, K. B., Mbaka, G. and Oluwe-Mimo, A. D. 2004. Histopathology of the

- gill and liver tissues of the African catfish *Clarias gariepinus* exposed to lead. *African Journal of Biotechnology*, 4(1): 117–122.
- Oyawoye, E. O. and Ogunkunle, M. 1998. Physiological and biochemical effects of raw jack beans on broiler. Proceedings of small conference of *Nigeria Society of Animal Production*, 23: 141–142.
- Perl, D.P. and Olanow, C.W. 2007. The neuropathology of manganese-induced Parkinson. *Journal of Neuropathology and Experimental Neurology*, 66(8): 675–682.
- Reish, D. L. and Oshida, P. S. 1987. Manual of methods in aquatic environmental research. Food and Agriculture Organisation (FAO), Rome, pp. 28.
- Robert, C. S. 2007. Water quality consideration for aquaculture. *Journal of Animal Ecology*, 2:1-15.
- Roels, H. A., Ghyselen, P and Buchet, J. P. 1992. Assessment of the permissible exposure level to manganese in workers exposed to manganese dioxide dust. *British Journal of Industrial Medicine*, 49: 25-34.
- Takeda, A. 2003. Manganese action in brain functions. *Brain Research Reviews*, 41, 79–87.
- Van Dyk, J. C. 2003. Histological changes in the liver of *Oreochromis mossambicus* (cichlidae) after exposure to cadmium and zinc. M.Sc. Thesis, Rand Afrikaans University, South Africa. pp. 28-62.
- W.H.O. (2008). UNICEF handbook on water quality. United Nations Children's Fund (UNICEF), New York. pp 45.