

GEOPHYSICAL AND HYDRO-CHEMICAL INVESTIGATION OF THE AREA AROUND A WASTE DUMP SITE IN ILE-IFE, SOUTHWEST NIGERIA

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ABSTRACT

An integrated geophysical and hydro-chemical investigation was carried out in the area around the Stadium-Apollo Market waste dump located in Igboya area of Ile-Ife. It was with the aim of determining the geoelectric characteristics and groundwater chemistry which were used in delineating the subsurface sequence and evaluating the possible pollution of groundwater in the area.

Eight Vertical Electrical Sounding (VES) stations were occupied along two traverse lines trending North-South and East-West directions. Dipole-Dipole profilings were also carried out along same traverses. Eight water samples were collected from hand dug wells and analyzed for some chemical elements such as total acidity, major ions (Na^+ , K^+ , Cl^- , Mg^{2+} , Ca^{2+}), NO_3^- and heavy metals (Cd, Cu, Fe, Hg, and Pb).

Results of the investigation indicate that a maximum of four subsurface layers- the top soil, clay/sandy clay unit, the weathered layer and the basement bedrock exist in the area. The weathered layer constitutes the main aquifer unit with thicknesses of between 7 and 38 m. From the 2-D resistivity structures, contaminant plumes characterized by relatively low resistivity values ($< 150 \text{ ohm-m}$) were identified down to depth of about 40 m in some places. Hydro-chemical results showed that the total acidity and concentration of Na^+ , Cl^- and heavy metals (Cd, Cu, Fe, Hg and Pb) in wells closest to and down slope of the waste dump (total acidity 5.1-10.0 mg/l; Na^+ 8.92-11.07 mg/l; Cl^- 127.62-297.78 mg/l; Cd, 0.006-0.009 mg/l; Cu, 0.330-1.840 mg/l; Fe, 2.300-4.600 mg/l; Hg, 1.247 mg/l and Pb, 0.260 mg/l) are higher than concentrations in distant wells, and most of them (Cl^- , Cd, Cu, Fe, Hg, and Pb) exceeded national and international drinking water quality standards.

It can be concluded that the groundwater in the area had been contaminated and the plume had migrated to a significant depth

KEYWORDS: Geophysical, Hydro-chemical, Waste Dump, Groundwater Quality, Pollution

INTRODUCTION

One of the common ways of waste disposal is by open dump. The primary consequence of solid waste disposal in open dump is the potential for gas and leachate generation due mainly to microbial decomposition of the waste materials. The migration of gas and leachate away from the waste dump and subsequent release into the surrounding environment, especially groundwater presents serious environmental problems including health hazard (Adepelumi et al., 2001, Tijani, et al., 2002, Samanjara and Bandara, 2003).

The Stadium-Apollo Market waste dump is located along Ife Township Stadium road off Obafemi Awolowo University Road 7 in Igboya area of Ile-Ife, Osun State. It is an active dump site that has been receiving a mix of solid wastes from within and around the neighborhood for over 30 years. The nature of waste mix disposed in the dump is heterogeneous and widely varied. Wastes such as garbage. (Waste fruits, meat, vegetables etc.); rubbish (paper, wood, scrap, rubber, plastic, nylon, leather, metals, glass, ceramics, aerosol cans and other house hold wastes); pathological waste

(human and animal wastes, dead animals, hospital waste etc.); industrial wastes such as chemicals, waste batteries from mechanic workshops, damaged electrical appliances, vehicles spare parts and agricultural wastes (farm animal manure, crop residue animal dung etc) are visible in the waste dump site.

The inhabitants of Igboya area and environs depend mainly on groundwater from wells to meet most of their daily water needs. Since these wells are generally shallow and located within the vicinity of the waste dump, there is the possibility that the groundwater in the area may have been contaminated through leachate percolation. Meanwhile, there have been reported cases of impure/coloured water from wells closest to the dump site. This study therefore intends to establish the extent of the contamination within and around the study area, identify the aquifer unit(s) and ascertain the presence of contaminant plumes within the subsurface. It also intends to determine specific contaminants through hydro-chemical analysis of sampled water and also determine the direction of contaminant migration.

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THE GEOLOGY OF THE AREA

The Ife-Ilesha area, which encompasses the present study area, is underlain by the Precambrian Basement Complex rocks of Southwestern Nigeria. Detailed descriptions of the geology of the area had been given by various workers including Elueze (1981), Ajayi (1981), Klemm et al (1988), Ige (1988) and Rahaman (1988).

According to Ige and Asubiojo (1991), rocks of Ife-Ilesha area include the gneisses, granites, amphibolites, schists and pegmatites (Fig 1). The local geology of the project site is schist. The rocks have generally undergone different degrees of weathering (Rahaman, 1998).

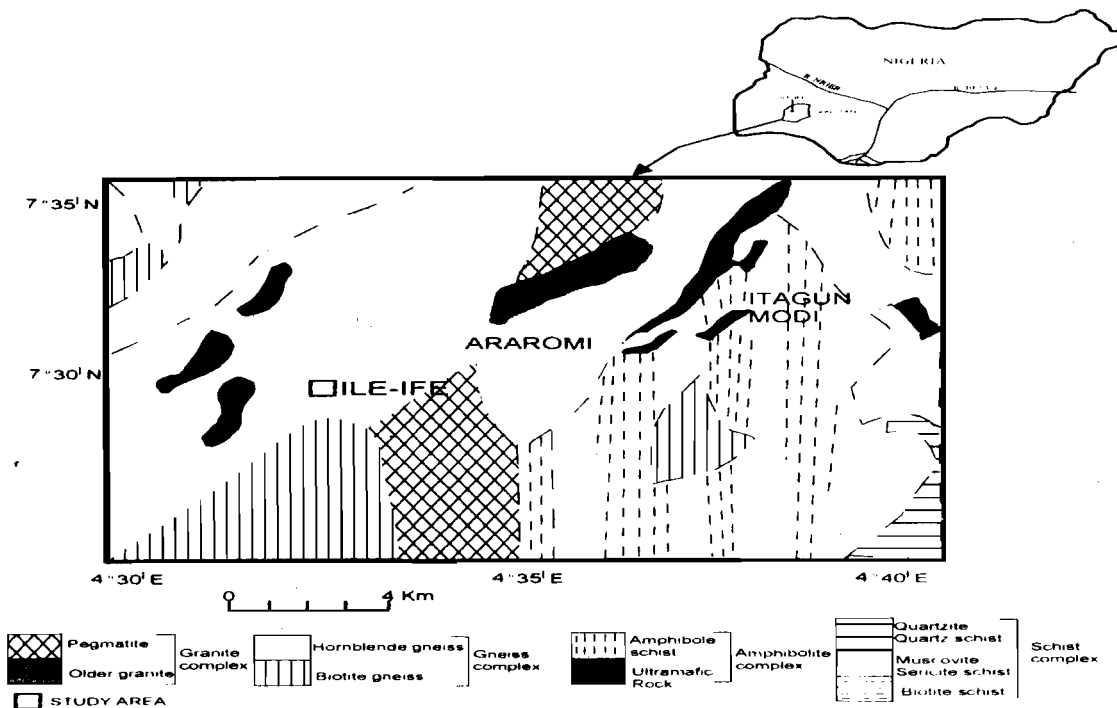


FIG. 1: GEOLOGICAL MAP OF THE AREA AROUND ILE-IFE (MODIFIED AFTER ELUEZE, 1981 AND IGE, 1988)

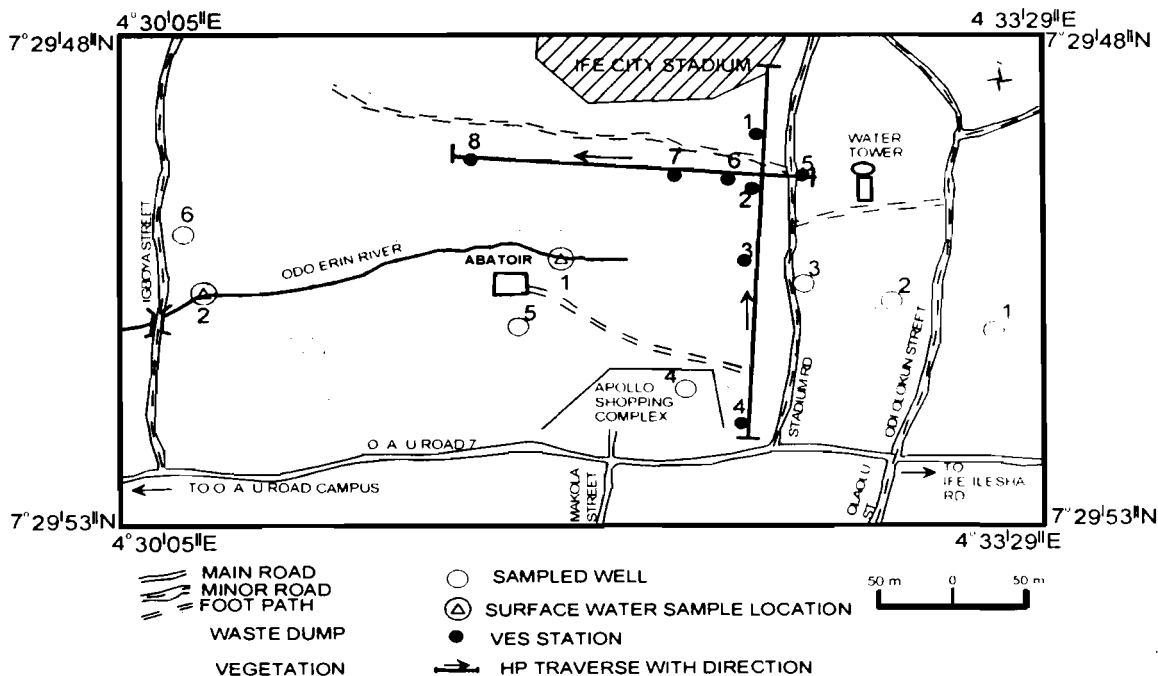


FIG. 2: MAP OF THE STUDY AREA SHOWING THE DUMP SITE

METHODS OF STUDY

The study involved geophysical and hydro-chemical investigations. The geophysical investigation involved the electrical resistivity method (Kearey and Brooks, 1988 and Bedient et al. 1998). The Vertical Electrical Sounding (VES) and the dipole-dipole profiling techniques were used in the study.

Eight Schlumberger VES stations were occupied along two traverses (Fig 2). The traverses trend approximately North-South and East-West. Measurements of ground resistance were made with the ABEM Terameter SAS 300C Resistivity Meter. The VES data were interpreted quantitatively by partial curve matching technique. The multi-layered field curves were interpreted segment-by-segment using theoretically generated master curves and associated auxiliary curves. The interpretation results from partial curve matching were used as started model parameters in 1-D forward modeling involving RESIST version 1.0 software. The final interpretation results were used to prepare geoelectric sections (Figs. 3 and 4).

Dipole-Dipole profiling was carried out along the two traverses. The inter-electrode spacing of 10 m was used while the expansion factor, n , was varied from 1-5. The dipole-dipole data were inverted to 2-D resistivity structures (Figs 5 and 6) using Diprho Version 4.01 software developed by Junk Ho Kim (1999-2001).

The hydro-chemical analysis involved the collection and analysis of water samples for chemical analysis. Eight water samples; three samples from wells in the up slope direction and three from wells in the down slope direction and the other two water samples were also collected from a nearby Odo Eran River. Sampling was done during the dry season. This was to ensure a true reflection of the quality of the groundwater, as there would be dilution due to rainfall or infiltration of surface run-off into the wells during the rainy season. The most remote wells (wells 1 and 6) (Fig 2) served as control wells. The study locations, including co-ordinates and water levels are presented in Table 1. The methods used for the analyses of the different parameters are summarized in Table 2.

Table 1: General Information on Water Sample Locations and Well Water Levels.

Sample Description	Longitude (E)	Latitude (N)	Elevation (m) a. s. l.	Distance from dump site	Static Water Level (m)
Well 1	004°33'33.333"	07°29'612"	286	150.0	8.2
Well 2	004°33'33.332"	07°29'659"	285	83.0	6.2
Well 3	004°33'325"	07°29'684"	280	19.0	5.4
Well 4	004°33'272"	07°29'688"	278	55.0	8.0
Well 5	004°33'225"	07°29'754"	270	183.0	7.0
Well 6	004°33'191"	07°29'891"	268	421.0	3.6
Odo Eran River location 1	004°33'239"	07°29'792"	267	141.0	-
Odo Eran River location 2	004°33'173"	07°29'859"	264	408.0	-

a. s. l.: Above Sea Level

Table 2: Methods used in Measuring the Different Chemical Parameters.

Parameters	Test methods
Total Acidity	KCL extraction / Titrimetric Method
Nitrate	Disulfuric Acid Method
Potassium ion	Flame Photometry Method
Sodium ion	Flame Photometry Method
Magnesium ion	Atomic Absorption Spectrophotometry Method
Calcium ion	Atomic Absorption Spectrophotometry Method
Chloride	Silver Nitrate Method
Heavy metals (Cd, Cu, Fe, Hg and Pb)	Atomic Absorption Spectrophotometry Method

Each determination was based on standard methods of the American Public Health Association (APHA, 1992); Chemical Methods of the British Drug House (BDH, 1978) and methods for physical and chemical analysis of fresh waters.

RESULTS AND DISCUSSIONS

Geoelectric Sections

The VES interpretation results and the type curves are shown in Table 3. The VES curves are the

HKH, HA, KH and QH type. The dominant KH type curve constitutes half of the VES curves (Table 3). Four subsurface layers were identified. These include the topsoil, clay/sandy clay, the weathered layer and the fresh basement (Figs. 3 and 4).

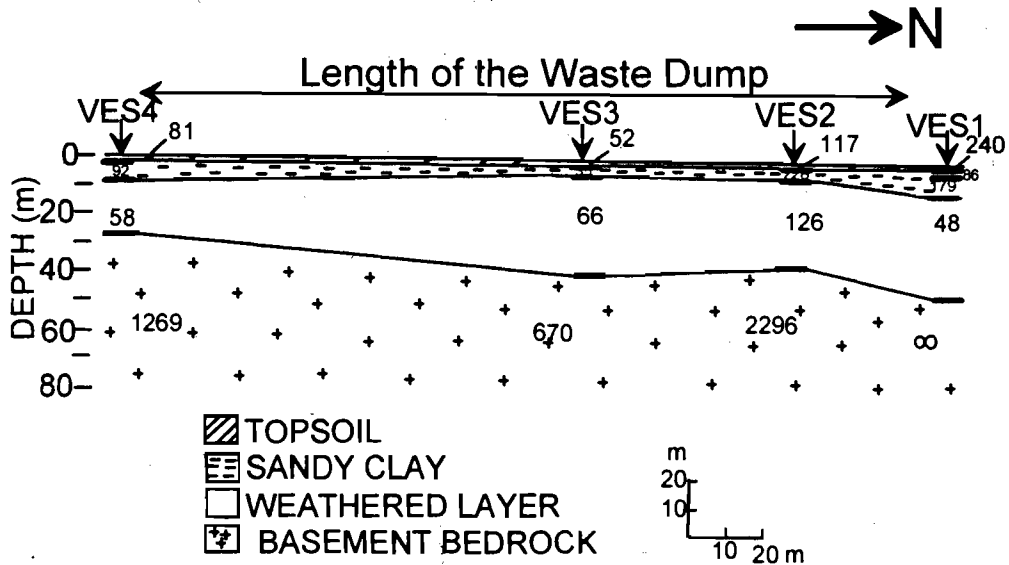


FIG. 3: GEOELECTRIC SECTION ALONG TRAVERSE 1: VES 1, 2, 3 AND 4

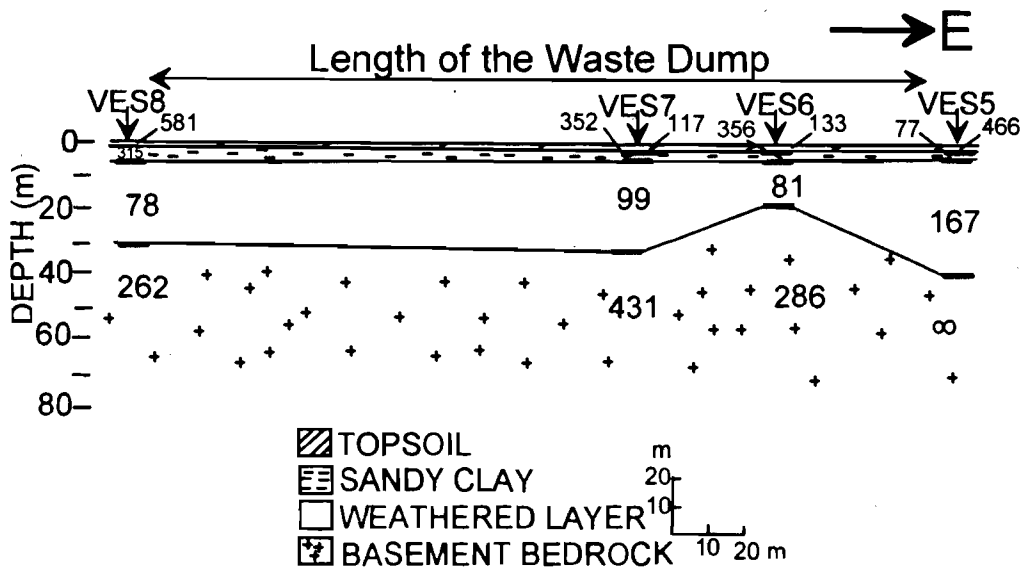


FIG. 4: GEOELECTRIC SECTION ALONG TRAVERSE 2: VES 5, 6, 7 AND 8

Table 3: VES Interpretation Results

VES Stations	Traverse Lines	Thickness (m) $h_1/h_2/h_3--/h_n$	Resistivity (ohm-m) $\rho_1/\rho_2/\rho_3---\rho_n$	Type curves
1	1	1.7/1.3/7.5/36.9	241/86/179/48/∞	HKH
2	1	0.6/4.2/4.0/30.2	117/228/126/2296	KH
3	1	0.8/2.2/36.1	52/35/66/670	HA
4	1	0.7/6.5/20.4	81/92/58/1269	KH
5	2	0.3/2.4/36.0	466/77/167/∞	HA
6	2	0.3/4.2/13.5	132/356/81/286	KH
7	2	0.3/3.5/27.9	117/352/99/431	KH
8	2	0.2/4.8/24.0	581/315/78/262	QH

The topsoil has layer resistivity and thickness ranging from 52-581 ohm-m and 0.2-1.7 m respectively. The 2.2-7.5 m thick clay/sandy clay layer has layer resistivity ranging from 35-365 ohm-m. The

weathered layer resistivities and thicknesses are in the range of 42-167 ohm-m and 13.5-36.5 m respectively. The basement bedrock has layer resistivity of 262-∞ ohm-m (Figs 3 and 4). Depth to the basement bedrock

varies from 27.6-46.4 m. The weathered basement constitutes the main aquifer unit. All hand dug wells around the study area exploit groundwater from this aquiferous unit.

Dipole – Dipole Sections

The 2-D resistivity structures obtained from the inversion of the dipole-dipole data are shown in Figures 5 and 6. The resistivity structure along Traverse 1 shows resistivity values that range from 14-3131 ohm-m, (Fig 5 (iii)). Low resistivity zones (< 150 ohm-m in green

colours) with significant depth extent of up to 40 m, suspected to be leachate impacted zones, are found between distances 50 and 80 m; 130 and 160 m and 185 and 200 m. These areas correlate with the actual location of the waste dump which lies between distances 47 and 206 m along the traverse (see Fig 2). The resistivity structure along Traverse 2 revealed marked variation in subsurface resistivity across the traverse. The resistivity values range from 19 –2364 ohm-m, (Fig 6 (iii)). The section shows areas with high and low resistivity values.

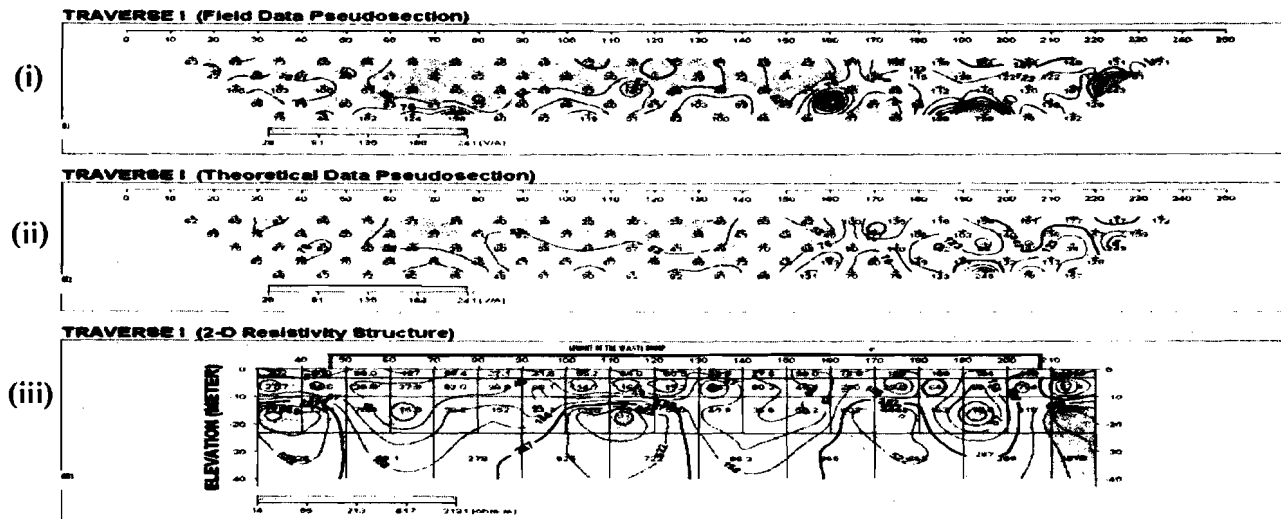


Fig 5: Dipole-Dipole Section along Traverse I – Stadium-Apollo Market Waste Dump Investigation (i) Field Data Pseudosection (ii) Theoretical Pseudosection (iii) 2-D Resistivity Structure

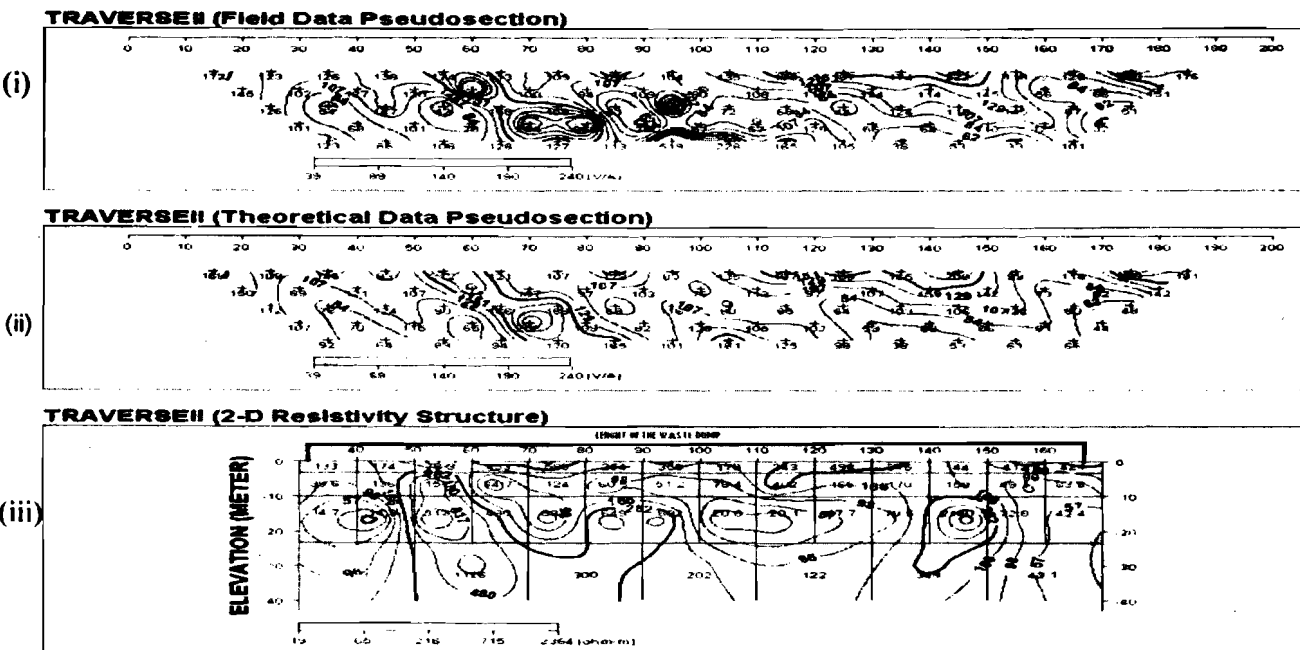


Fig 6: Dipole-dipole Section along Traverse II– Stadium-Apollo Market Waste Dump Investigation (i) Field Data Pseudosection (ii) Theoretical Pseudosection (iii) 2-D Resistivity Structure.

Areas with low resistivity values (< 150 ohm-m in green colours) are indicative of a conducting medium suspected to have been impacted. The low resistivity zones are between distances 1-46 m; 60-135 m and

150-170 m. The depth extent varies from 20 m to more than 40 m, particularly between distances 0 and 45 m; 110 and 140 m; and 150 and 170 m where leachate from the waste dump is suspected.

Based on field observations and 2-D resistivity structures obtained from the inverted dipole-dipole data, it can be inferred that:

- i. The subsurface in the study area is suspected to have been impacted by leachate from the waste dump;
- ii. Suspected impacted zones are characterized by relatively low resistivity (< 150 ohm-m) and
- iii. There has been a vertical movement of the leachate down to depths of about 40 m.

Hydro-Chemical Analysis

Results of hydro-chemical analysis are presented in Tables 4 and 5. The results show that wells closest to the waste dump and those located down slope of the waste dump have higher concentrations of the analyzed parameters.

Total acidity values were highest (10 mg/l) in well 3 (20 m from the dump). Well 6 which is located 430 m from the waste dump has the lowest value (2.5 mg/l) of total acidity (Table 4). The results of total acidity showed that the closer the well to the dump, the higher is the value of total acidity. Waste batteries from mechanic workshops and from household that are disposed in the waste dump are suspected to be the source of total acidity in the groundwater of the area.

Na⁺, like total acidity, showed a gradual decrease in concentration with increasing distance of sampled wells away from the waste dump (Table 4).

Wells 3 (20 m up slope) and 4 (55 m down slope) are the closest to the dump and their concentration values for Na⁺ are 10.7 mg/l and 11.07 mg/l respectively (Table 4). The least values of Na⁺ (3.01 and 2.85 mg/l) were detected in wells 1 and 6 respectively. These are control wells located further away (160 m and 430 m) from the. Human and animal wastes and decomposition of various minerals are the major contributors of Na⁺ in groundwater (USGS, 2005).

The concentrations of K⁺ were higher in wells 4 and 5 (3.05 mg/l and 2.63 mg/l respectively), (Table 4). The concentrations of K⁺ in other wells are low (1.58-2.01 mg/l) and could represent background values. Usually, K⁺ in the environment is derived from potash salts and mines (Scoth and Smith, 1981).

The concentrations of NO₃⁻ (Table 4) are higher in wells 3 (1.80 mg/l), well 6 (1.02 mg/l) and in surface water samples 1 and 2 (141 m and 403 m respectively down slope), compared to values detected in other wells (Table 4). The lowest values of NO₃⁻ (0.20-0.60 mg/l) are detected in wells 1, 2, 4 and 5. A NO₃⁻ concentration of less than 0.2 mg/l represents natural conditions while values greater than 3 mg/l may indicate human effects. NO₃⁻ concentration has not exceeded the maximum contaminant level (m. c. l.) set by WHO (2004), FEPA (1991) and USGS (2002). Animal and human wastes, sewage systems garden fertilizer and livestock facilities are major contributors of NO₃⁻ to groundwater but the largest share comes from society's wastes.

Table 4: Total Acidity, Na⁺, K⁺, Cl⁻, NO₃⁻ and Total Hardness in the Water

Up slope	Sample description	Total Acidity (mg/l)	Na ⁺ (mg/l)	K ⁺ (mg/l)	Cl ⁻ (mg/l)	NO ₃ ⁻ (mg/l)	Total Hardness (mg/l)	Distance from dump site (m)
	Well 1	3.0	3.01	1.58	113.44	0.20	49.7	160.0
	Well 2	5.0	8.92	2.01	184.34	0.22	54.8	82.5
	Well 3	10.0	10.71	1.69	297.78	1.80	50.4	20.0
Waste dump								
Down slope	Well 4	3.4	11.07	3.05	127.62	0.26	61.7	55.0
	Well 5	5.1	9.06	2.63	255.24	0.60	18.1	180.0
	Well 6	2.5	2.85	1.72	113.44	1.02	16.2	430.0
	Surface water location 1	4.4	3.72	1.88	85.08	1.70	25.7	141.0
	Surface water location 2	3.4	4.05	1.99	170.16	1.89	19.7	403.0

Groundwater in the study area belongs to the soft water category (0-60 mg/l of CaCO₃). However, well 4, located 55 m down slope from the waste dump and within the Apollo Market Complex displays moderate hardness (61-120 mg/l of CaCO₃). Hardness of water results from dissolution of metallic ions in water, dissolved limestone and decaying fruits and other calcium rich vegetable matter.

Considering Cl⁻, the concentrations detected were highest in wells 3 (20 m upslope) and 5 (180 m down slope). Table 4 shows that Cl⁻ concentrations increase with decreasing distance from the waste dump in the up slope direction. Cl⁻ values in well 3 (297 mg/l) and well 5 (255 mg/l) exceeded the WHO, FEPA and

USEPA standard of 200 mg/l. Cl⁻ is usually found in close association with Na⁺. It is leached into the environment from such household wastes as refrigerators, air conditioners and plastic foams. These materials contain CF₂Cl₂ and CFC₃, which are used as propellants in the manufactured of these products.

Cd displays a gradual increase in concentration as you move closer to the waste dump. The highest concentration value of 0.009 mg/l was detected in well 5, down slope (Table 5). Well 4 with a concentration value of 0.004 mg/l exceeded the maximum contaminant level (m. c. l.) of 0.003 mg/l set by the World Health Organization but less than the USEPA standard of 0.005 mg/l. In well 3 and 5 (Table 5), the concentration

exceeded both the WHO and USEPA standards but less than the FEPA standards of 0.010mg/l. Incinerator ash, landfill leachate, metal plating and erosion of battery wastes contribute heavily to Cd load in groundwater of the area.

Considering Cu, well 3 has the highest concentration value of 1.840 mg/l (Table 5). The value

exceeded the maximum contaminant level (m. c. l.) set by FEPA (1991). Wells 4 and 5 did not show the presence of Cu. Generally, corrosion of household plumbing systems, industrial and domestic wastes release Cu into the environment and groundwater. Household wastes also contain Cu (from copper plated materials, cooking vessels and most electrical items) that ultimately enters into groundwater.

Table 5: Heavy Metal Concentrations in Water Samples From the Study Area

Up slope	Sample Description	Cd (mg/l)	Cu (mg/l)	Fe (mg/l)	Hg (mg/l)	Pb (mg/l)	Distance from dump site (m)
	Well 1	0.000	0.050	2.300	0.000	0.000	160.0
	Well 2	0.001	0.330	0.000	0.000	0.020	82.5
	Well 3	0.006	1.840	2.300	1.247	0.260	20.0
Waste dump							
Down slope	Well 4	0.004	0.000	3.200	0.000	0.000	55.0
	Well 5	0.009	0.000	4.600	0.000	0.000	180.0
	Well 6	0.003	0.860	2.600	0.000	0.000	430.0
	Surface water location 1	0.002	1.380	0.700	0.000	0.000	141.0
	Surface water location 2	0.000	1.410	1.000	0.000	0.000	403.0

The concentrations of Fe increase with decreasing distance from the waste dump. Besides well 2, all other sample stations indicated the presence of Fe (Table 5). The concentration of Fe in groundwater of the study area exceeded drinking water quality standards of 1.00 mg/l and 0.300 mg/l set by FEPA (1991) and USEPA (2005) respectively. Wells 4 and 5 recorded the highest concentrations (Table 5). The high Fe content indicates that leachate from the waste dump is rich in Fe, probably from scrap metals disposed in the dump site.

Considering Hg and Pb, only well 3 indicated the presence of Hg while Pb was detected in wells 2 and 3 (Table 5). The concentration of Hg and Pb exceeded drinking water quality standard of .0001 mg/l for Hg and 0.050 mg/l for Pb (FEPA, 1991); 0.001 mg/l for Hg and 0.010 mg/l for Pb (WHO, 2004) and 0.002 mg/l for Hg and 0.015 mg/l for Pb (USEPA, 2005). Hg can get into groundwater from pesticides, electrical equipment (e.g. batteries, lamps and switches), run-offs from landfills and croplands. Pb can get into leachate and subsequently groundwater from corrosion of household plumbing systems, pesticides, household paints, leachate from waste batteries, pottery, cooking vessels, landfill leachate and incinerator ash.

CONCLUSIONS

The geoelectric sections identified a maximum of four subsurface layers namely, the topsoil, the sandy clay/clayey sand unit, the weathered basement layer and the fresh basement unit. The weathered layer constitutes the main aquifer unit. This result is in agreement with earlier reports by Okhue and Olorunfemi (1991) and Olorunfemi and Okhue (1992).

Dipole-dipole resistivity structures revealed that the subsurface layers have been contaminated by leachate from the waste dump. This is indicated by the presence of contaminant plumes represented by the typically low resistivity values (< 150 ohm-m) and significant depth of over 40 m in some places (Figs. 5 and 6).

Hydro-chemical results confirmed the suspected subsurface and groundwater contamination from geophysical investigation results. The relatively high concentrations of total acidity and Na⁺ suggest contamination from the waste dump. Cl⁻ concentrations from wells 3 and 5 (297.78 mg/l and 255.24 mg/l respectively) exceeded the FEPA USEPA, and WHO water quality standards of 1991, 2002 and 2004 respectively. Concentrations of Cl⁻ in other wells are also relatively high. All the wells sampled indicated the presence of heavy metals (Cd, Cu, Fe, Hg and Pb). Fe (2.300 mg/l wells 1 and 3; 3.200 mg/l well 4 and 4.600 mg/l well 5), Hg, (1.247 mg/l well 3) and Pb (0.020 mg/l and 0.260 mg/l well 3) concentrations exceeded FEPA, USEPA and WHO water quality standards. Cd values in wells 3 and 5 (0.006 mg/l and 0.009 mg/l respectively) exceeded USEPA and WHO standards of 0.003 mg/l and 0.005 mg/l respectively while Cu concentrations in wells 2, 3, 6 (0.330 mg/l, 1.840 mg/l, and 0.860 mg/l respectively) and from surface water samples for locations 1 and 2 (1.380 mg/l and 1.410 mg/l respectively) exceeded the FEPA standard of 0.100 mg/l.

Generally, the presence, at relatively high concentrations, of environmental tracers (major ions, particularly Na⁺ and Cl⁻) indicates that the groundwater contamination in the area is as a result of human activities.

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