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BIOSORPTION OF IRON BY HEAVY-METAL TOLERANT *Micrococcus* sp.

Ladifa H. Muhammad and Muhammad B. Ibrahim

Department of pure and industrial chemistry, Bayero University, Kano.

Email: ladifa_mansur@yahoo.com; mbibrahim.chmbuk@edu.ng.

ABSTRACT

The presence of metal ions in aqueous solutions represents, nowadays a major environmental problem. The present study investigates the potential adsorption capacity of *Micrococcus* sp. for the removal of iron (III) ions from aqueous solution. The adsorbent was characterized using FTIR analysis, in order to reveal the presence of functional groups. Batch adsorption experiments were carried out to evaluate the effect of process parameters such as pH, contact time, initial iron concentration, temperature and adsorbent dosage on removal capacity. The optimum removal of iron (III) ions was obtained at pH 6 with 94.2%, adsorbent dose of 0.5g with 92.5% and equilibrium time of 30 min with 88.6%. Kinetic data were successfully described with pseudo-first order, pseudo-second order and elovich models. The results show that the biosorption process of Fe (III) followed pseudo second order with $R^2 = 0.999$. The equilibrium data was further tested using Langmuir, Freundlich and Temkin models, and was found to follow the Langmuir isotherm ($R^2 = 0.978$). Thermodynamic parameters such as ΔG° , ΔH° and ΔS° were evaluated and it has been found that the biosorption process was feasible, spontaneous and exothermic in nature.

Keywords: Bacteria, Biosorption, Iron, Isotherm, Kinetic

INTRODUCTION

Heavy metal pollution occurs directly by effluent outfall from industries, refineries and waste treatment plants and indirectly by the contaminants that enter the water supply from soil/ground water systems and from the atmosphere via rain water. Among toxic substance reaching hazardous levels are heavy metals (Vieira & Volesky, 2000). Heavy metals are the group of contaminants of concern, which comes under the inorganic division. Exposure to heavy metals has been linked with developmental retardation, various cancers, kidney damage, and even death in some instances of exposure to very high concentrations (Ahmaruzzaman, 2011; Gawali *et al.*, 2014). Heavy metals do not degrade into harmless end products and are generally accumulated along the food chain (water–plants–animals–humans), thus, posing a great threat to living organisms (Benaisa *et al.*, 2016).

Metals play an integral role in the life processes of living organisms. Heavy metals are metals with densities higher than 5 g/cm³. Some metals (Ca, Co, Cr, Cu, Fe, K, Mg, Mn, Na, Ni and Zn) are essential, they serve as micronutrients and are used for redox-processes, to stabilize molecules through electrostatic interactions; as components of various enzymes; and regulation of osmotic pressure. While many other metals (e.g., Ag, Al, Cd, Au, Pb, and Hg) have no biological role and they are nonessential. They are potentially toxic to living organisms, especially microorganisms. Toxicity of nonessential metals occurs through the displacement of essential metals from their native binding sites or through ligand interactions. Heavy metals in wastewater come from industries and municipal sewage, and they are one of the main causes of water and soil pollution (Ashok *et al.*, 2010). These practices leave persistent toxic heavy metals like chromium, nickel, lead, zinc,

cadmium and copper, which tend to accumulate and deteriorate the environment (Abbas *et al.*, 2010). Contamination of heavy metals in the environment is a major concern because of their toxicity and threat to human life and environment (Rajendran & Gunasekaran, 2007). These heavy metal ions are toxic to both human beings and animals. The toxic metals cause physical discomfort and sometimes life threatening illness and irreversible damage to vital body system (Malik, 2004).

In another different approach, Zully (2012), studied biosorption of Lead and Copper by heavy metal tolerant *Micrococcus luteus* DE2008. They find out that Lead had no effect on the biomass expressed as mg carbon cm³ of *Micrococcus luteus* DE2008, but in the case of Copper, the minimum metal concentration that affects the biomass was 0.1mM Cu (II). According to these result this microorganisms shows a greater tolerance for Lead.

Abioye *et al.* (2015) study biosorption of zinc and manganese by *micrococcus variance* and *staplylococcus Aureus* isolated from agricultural soil. In their studies, they investigated that the metal concentrations were determined using an Atomic Absorption spectrophotometric (AAS).

Iron as a common contaminant enters the water bodies through effluents of industries such as electroplating, mining and steel processing among others (Pavani & Kumar, 2013). The presence of Fe in water results in undesirable colour, odour and taste which makes water unfit for industrial and domestic consumption. Exceeding iron concentration greater than 0.3 mg/L causes water staining that adversely affect plumbing fixtures, dishware and clothes and produce a yellow to reddish appearance in water. These levels may also impart taste and odor to drinking water (Shokoohi *et al.*, 2009).

Many Methods exist for removal of heavy metals from industrial wastewater, these methods includes: Physico-chemical method, Chemical Precipitation, Coagulation and Flocculation, Electrochemical Treatments, Ion Exchange, Membrane Filtration, Biological Methods, Electrodialysis(Gunatilake, 2015). The aim of the present work is to investigate iron biosorption using dead *Micrococcus sp.*

MATERIALS AND METHODS

Sample Collection and Isolation of the Bacteria

The effluent sample was collected from Jana Plastic Industry Sharada phase II. The collected sample was transferred to sterile glass bottle container and immediately taken to the laboratory for isolation, in which 1ml of effluent was serially diluted using 10ml of distilled water and spread over nutrient agar solid plates. The plates were then incubated at $35^{\circ}C$ for 24hrs.

Selection of Heavy Metal Tolerant Bacterial Strain

Iron resistant bacterial strain was isolated from the effluent using bacterial nutrient agar and broth medium, which was amended with 0ppm, 50ppm, 100ppm and 200ppm of Fe and standard spread plate method was performed. The inoculated plate was incubated at $35^{\circ}C$ for 24hrs. After 24hrs incubation larger identical colonies from each plate were isolated. These isolates were characterized and further employed for the iron removal. Morphological, physiological and biochemical characteristics of the isolated bacteria species were carried out (Christopher *et al.*, 2014).

Bacterial Biomass Production

Selected bacterial strain was maintained on nutrient agar medium. Biomass of bacteria was produced by growing bacterial culture in nutrient broth in a conical flask at $35^{\circ}C$ for 24hrs. Cells were harvested by centrifugation at 7500rpm for 20min. Cell pellet were washed three times with distilled water, autoclaved at $121^{\circ}C$ for 15min to kill the bacteria, After that the cell was dried at $100^{\circ}C$ for 48hrs in an oven (Mohammed *et al.*, 2013).

FTIR Analysis

FTIR Analysis was used to identify the chemical functional groups present in the fresh biomass and Fe-loaded biomass using $650-4000\text{cm}^{-1}$ Fourier transform infrared spectroscopy (Cary 630; Agilent Technologies).

Batch Biosorption Experiments

The biosorption of iron (III) on the obtained biomass was performed in a batch system. In the present work several effects on the biosorption of iron were

studied, such as the initial pH, the contact time and kinetics, equilibrium parameters and thermodynamics. The effect of pH was studied by performing series of experiments at a pH ranging from 2.0 to 10, solutions of 0.1M HCl and 0.1 M NaOH were used to adjust the adsorbate pH at room temperature. The kinetic study of the biosorption of Fe (III) was carried out in glass bottles containing 100 ml of Fe (III) with a concentration of 10 mg /L mixed with 0.5 g of dried biomass. The pH was adjusted to the optimum value and the bottles remain in agitation at 150 rpm for 30min at $30^{\circ}C$. Aliquots of 2 ml were taken every 5 min for the first 20 minutes then every 10 min, and then every 20 min for the rest of the experiment. Next the samples were filtered and analyzed by measuring the absorbance using AAS in order determine the residual concentration of iron. The equilibrium study of Fe (III) removal was performed in glass bottles containing different concentrations of iron (10, 20, 30,40 and 100 mg/L) and 0.5 g of dried biomass. The bottles are placed in agitation (150 rpm) for 30min at room temperature. The initial pH is adjusted as described above. The thermodynamic of Fe (III) removal was performed in glass bottleshaken at $25^{\circ}C$, $30^{\circ}C$, $35^{\circ}C$ and $40^{\circ}C$, and 150rpm on orbital shaker. Centrifugation was carried out after the defined contact time and supernatant was carried for analysis using Atomic adsorption spectrophotometer AAS (Perkin Elmer, Analyst 200).

RESULTS AND DISCUSSION

FTIR spectral analysis

Biosorption in bacteria may mainly be attributed to the cell wall structure containing functional groups such as amino, hydroxyl, and organic phosphate group ((Wang, 2007; Magdalena & Małgorzata, 2014)). The FT-IRanalysis was performed to obtain information about the interaction between the functional groups of biomass and Fe (III) ions. The FT-IR spectra of the bacterial biomass, before and after iron treatment are shown in Figure 1 and 2. The vibration around 3342cm^{-1} - 3316cm^{-1} is indicative of the presence of the -OH group on the bacterial biomass. The peak around 2959cm^{-1} could be assigned to the C-H stretch. By comparing iron loaded biomass with FTIR spectra of pure biomass significant the changes in some peaks were observed. As can be seen in Figure.2, the peak at 3335cm^{-1} - suggests the major role of -OH and 1598cm^{-1} -NH group for Fe (III) biosorption onto *Micrococcus sp.*.These observations indicate that (hydroxyl and amine) functional groups on the surface of the biosorbent are responsible for the iron binding which is a strong bond.

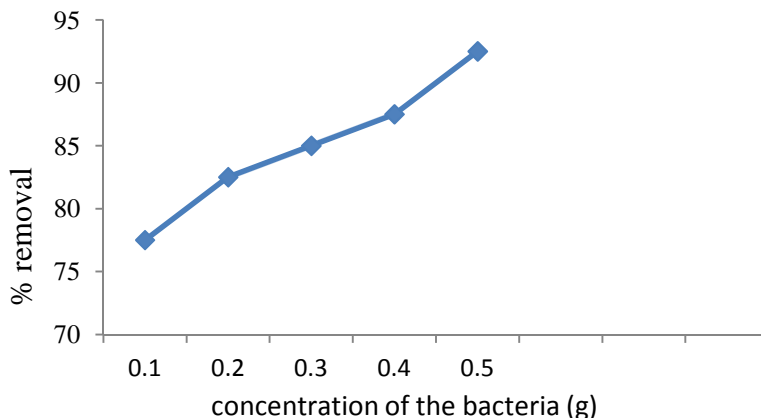


Figure 4: Effect of Biosorbent Dosage on % Removal

Effect of Initial Concentration

The biosorption experiment was conducted at different initial concentrations (10, 20, 30, 40 and 100mg/L) using an optimum time of 30min and a close of 0.5g of the bacterial cell. The percentage removal of iron from aqueous solution decrease with increase in the initial concentration from 92.5% to 40% while amount absorbed q_e values increases with increase in the initial concentration. This variation is represented in Figure (5), which data revealed that, the biosorption potential was higher at lower initial concentration of iron added to aqueous solution of the bacterial cell. These characteristics indicated that the surface saturation of biosorbent depends on the initial

concentration. The greater uptake of metal by the biosorbent materials at the lowest concentration could probably be due to a rapid metal absorbing ability of bacterial cell. The results are in good agreement with that of Christopher *et al.*, (2014) and Mohammad *et al.*, (2013) during the biosorption of heavy metals by *Bacillus thuringiensis* strain OSM29 originating from industrial effluent contaminated north indian soil. Similarly this is in agreement with study of Harikrishna & Nanganuru (2012) during the biosorption of cadmium by *pseudomonas putida*, who mentioned that the maximum initial metal concentration of Cd was obtained at 10mg/L .

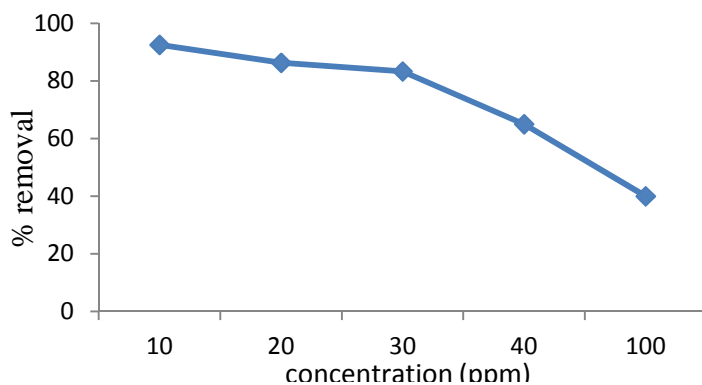


Figure 5: Effect of Initial Concentration on % Removal

Effect of pH

pH is an important parameter influencing heavy metal biosorption from aqueous solutions. It influences not only the surface charge of the adsorbent, the degree of ionization of material present in the solution, but also the dissociation of functional groups on the active sites of the adsorbent (Panida *et al.*, 2014). The effect of pH was studied in range 2,4,6,8 and 10 as shown in Figure (6). The percentage removal of Fe was observed to depend on the pH, the percentage removal increases with pH from 2.0 to 6.0 for both dead and living bacterial cells, and above these levels, a further increase in pH decreases the percentage

removal. At low pH values, there is a competition between the protons Fe ions on the sorption sites. At higher pH, in soluble iron hydroxide starts to precipitate from the solutions. The maximum removal efficiency was 94.2 % and 90.4 % at pH 6.0. This result agrees with the study of Mohammad *et al.*, (2013) in which the optimum biosorption pH value of Cd^{2+} by *Bacillus thuringiensis* strain OSM29 was 6.0. Similarly this result is in agreement with the result obtained by Guven *et al.*, (2004) in which the optimum adsorption pH values of cadmium(II) was determined as 6.0.

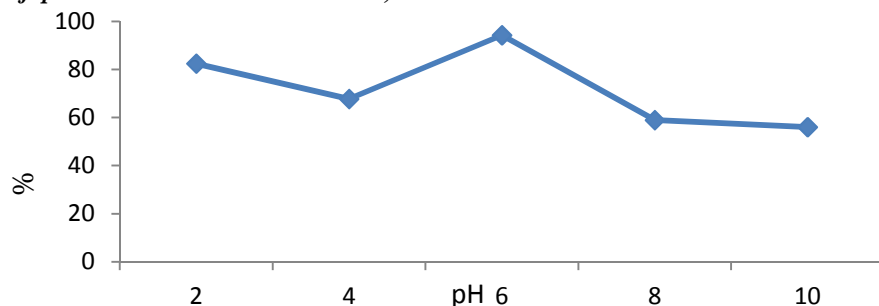


Figure 6: Effect of pH on % Removal

Effect of Temperature

In this research, biosorption of Iron by *Micrococcus* sp. appears to be temperature dependent. Maximum removal of iron was 30°C with the percentage removal of 93.1% as shown in Figure (7), which may depend on cell metabolism that are mostly likely to be inhibited by low temperature, the higher temperature also affect the integrity of the cell membrane and also affect the cell wall configuration. This is in agreement

with the result obtained by Ritixa & Monika (2013) in which the optimum temperature for Fe was 30 °C with percentage removal of 92% during the study of the effect of pH and temperature on the biosorption of heavy metals by *Bacillus licheniformis* and similarly this is in agreement with study of Abd El-Raheem *et al.*, (2013) who mentioned that the maximum biosorption rate for *Cd*, *Co*, *Zn* and *Pb* by *Enterobacter* sp. Was obtained at 30°C.

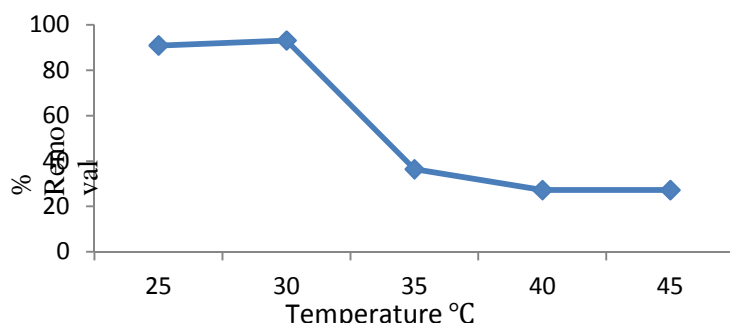


Figure 7: Effect of Temperature

Biosorption Kinetics

Different kinetic models including pseudo-first-order, pseudo-second-order and Elovich model were tested, in order to study the mechanism through which the biosorption process.

Pseudo-first-order equation

Pseudo first order integrated rate equation can be represented by the linear equation:

$$\log (q_e - q_t) = \log q_e - \left(\frac{k_1}{2.303}\right) t \quad (1)$$

The experimental results of the first order rate constants are presented in Table 1. The biosorption data have a low regression coefficient which suggests that the adsorption of iron on the biosorbent does not follows a pseudo-first order adsorption kinetics.

Pseudo-second order equations

Pseudo second order integrated rate equation is defined by the relation:

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} t \quad (2)$$

Values of k_2 and equilibrium adsorption capacity (q_e) were calculated from the intercept and slope of the plots of $\frac{t}{q_t}$ vs. t , respectively, the values were

presented in Table 1, From which , the correlation coefficient of Fe^{3+} was 0.995 which is better than that of both the pseudo- first and Elovich models. This result agreed with that reported by Arezoo *et al.*, 2015, in which the same conclusion indicated that the second-order equation correlates well with the adsorption studies.

Elovich equation

The Elovich equation based on the adsorption capacity and the rate expression can be presented in a linear form as:

$$q_t = \frac{1}{\beta} \ln (\alpha \beta) + \frac{1}{\beta} \ln (t) \quad (3)$$

The constants α and β were obtained from the slope and intercept of the linear plot of qt versus $\ln t$ as shown in Table 1. The data did not show a good correlation with the Elovich equation. The correlation coefficient for the linear plot, R^2 , suggests a poor relationship between the parameters and also explained that the process of biosorption do not fit Elovich model.

Table 1: Kinetics parameters for biosorption of Iron

Metal	Pseudo first order			Pseudo second order			Elovich equation		
	$k_1(\text{min}^{-1})$	$q_e(\text{mg/g})$	R^2	k_2	q_e	R^2	α	β	R^2
Fe	-0.034	0.904	0.189	0.117	9.183	0.999	1.188	2.426	0.864

Langmuir Isotherm

The Langmuir isotherm model is based on the assumption that there is a finite number of active sites which are homogeneously distributed over the surface of the adsorbent. These active sites have the same affinity for adsorption at a mono molecular layer level and there is no interaction between adsorbed molecules. A well-known linear form of the Langmuir equation can be expressed as in eqn (4):

$$\frac{C_e}{q_e} = \frac{1}{K_L Q_m} + \frac{C_e}{Q_m}$$

where: K_L is the Langmuir adsorption constant (Lmg^{-1}), and Q_m is the theoretical maximum (monolayer) adsorption capacity (mgL^{-1}) (Boparai, 2011). A plot of C_e/q_e against C_e is a straight line graph with $1/Q_m$ as the slope and $\frac{1}{K_L Q_m}$ as the intercept. The values of Q_m and K_L constants with the correlation coefficients for Langmuir isotherm are presented in Table 2. The maximum monolayer capacity (Q_m) was found to be 6.36mgg^{-1} . The correlation coefficient was extremely high (> 0.9). The biosorption process is found to be favorable since the R_L values lies between 0 and 1.

Freundlich Isotherm

The Freundlich isotherm model can be applied to a adsorption on a heterogeneous surface between the adsorbed. The Freundlich isotherm is given in eqn 5:

$$\log q_e = \log K_f + \frac{1}{n_f} \log C_e$$

Where K_f is the Freundlich constant and $(1/n_f)$ is the heterogeneity factor (Boparai, 2011). Table 2 shows the Freundlich adsorption isotherm constants and the respective correlation coefficients $\frac{1}{n_f}$ and K_f constants were determined from the slope and intercept of the plot of $\log q_e$ against $\log C_e$. The value of $1/n_f$ (heterogeneity factor) was between 0 and 1. For heterogeneous surface, the $\frac{1}{n_f}$ value is close to zero.

The data in table 2 shows that for both the metals, $1/n_f$ values were less than one which shows that n_f is greater than one, indicating that the biosorption of iron onto bacterial cells are favorable physisorption (multilayer) biosorption process.

Temkin Isotherm

The Temkins isotherm model was also applied to the experimental data, unlike the Langmuir and Freundlich isotherm models; this isotherm takes into account the interactions between adsorbents and adsorbate and is based on the fact that the free energy of adsorption is simply a function of surface coverage (Surchi, 2011; Tan, et al., 2008). The linear form of the Temkins isotherm model equation is given in eqn 6

$$q_e = B_T \ln K_T + B_T \ln C_e \tag{6}$$

$$B_T = \frac{RT}{b_T}$$

T is the absolute temperature in Kelvin, R is the universal gas constant ($8.314 \text{Jmol}^{-1}\text{K}^{-1}$), and B_T is related to heat of adsorption, while K_T is the equilibrium binding constant (Lmg^{-1}). This isotherm was applied by a linear plot of q_e against $\ln C_e$, the constants B_T and K_T were calculated from the slope and intercept respectively, and are presented in Table 2 with the R^2 value. Again, looking at the regression R^2 obtained (0.726), it is seen that this isotherm is not applicable for the description of equilibrium data. However, comparing the isotherms applied, the Langmuir isotherms gave the best fit, followed by the Freundlich isotherm and then the least fit was obtained with the Temkin isotherms.

Table 2: Biosorption Isotherm parameters

Metal	Langmuir isotherm				Freundlich isotherm			Temkin Isotherm		
	Q_m	K_L	R_L	R^2	K_f	$1/n$	R^2	K_T	B_T	R^2
Fe	6.361	0.164	0.058	0.979	2.013	0.252	0.970	8.443	0.869	0.726

Biosorption thermodynamics

Thermodynamic parameters such as the free energy standard free energy change (ΔG^0), enthalpy change

(ΔH^0) and entropy change (ΔS^0) changes during adsorption can be evaluated from the following Equations (7).

$$K = \frac{C_{ads}}{C_e}$$

$$\Delta G^0 = -RT \ln K \tag{7} \quad \text{and}$$

$$\ln K = \frac{\Delta S^0}{R} - \frac{\Delta H}{RT}$$

Where K is the equilibrium constant, T is the absolute temperature in Kelvin, and R is the universal gas constant (8.314Jmol⁻¹K⁻¹). The ΔH° and ΔS° values were obtained from the slope and intercept of the Vant Hoff's plots (lnK vs. 1/T) for the metals adsorption from aqueous solution onto the microbial cells at various temperatures. The values obtained are tabulated in Table 3. The values of the enthalpy change (ΔH°) and the entropy change (ΔS°) found in this work were -152.3kJ/mol and -0.485kJ/mol/K,

respectively. The negative ΔG° value indicates that the process is feasible and spontaneous in nature. The negative ΔH° value indicates the exothermic nature of biosorption and the negative value ΔS° corresponds to a decrease in degree of freedom of the adsorbed species. This is in agreement with result obtained by Senthil *et al* (2010) in which ΔG°, ΔH° and ΔS° are all negative during the thermodynamic and kinetic studies of cadmium adsorption from aqueous solution onto rice husk.

Table 3: Thermodynamic parameters

Metal	Temperature (k)	ΔG° (KJ/mol)	ΔH° (KJ/mol)	ΔS°(kJ/mol.K)
Fe	298	-7.90	-152.3	-0.485
	303	-5.48		
	308	-3.06		
	313	-0.63		

CONCLUSION

From the present study it could be concluded that bacteria play a very important role in the removal of heavy metals from waste water. The bacteria (*Micrococcus* sp.) were isolated from industrial effluent collected from Jana plastic industry Sharada using nutrient agar (NA) and broth media,Langmuir isotherm fitted well with the biosorption process. Biosorption is moreover influenced by various parameters such as time, dosage, initial metal concentration, pH, and temperature.The kinetic data were analyzed using pseudo-first order, pseudo-

second order and elovich equation. From the linear regression coefficient values the data were found to be best fitted to pseudo second order kinetics.The thermodynamic parameters illustrated the biosorption process is feasible, spontaneous and exothermic

Conflict of interest

There is no conflict of interest.

Contribution of individual authors

Ladifa Hamisu Muhammad was the student who performed the experiments. Prof. Muhammad B. Ibrahim was her thesis supervisor.

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