

Full Length Research Paper

Production and characterization of heavy-metal removing bacterial biofloculants

J. Lin* and C. Harichund

School of School of Life Sciences, University of KwaZulu-Natal (Westville)
Private Bag X 54001, Durban, Republic of South Africa.

Accepted 5 September, 2011

Biofloculants produced by *Herbaspirillum* sp. CH7, *Paenibacillus* sp. CH11, *Bacillus* sp. CH15 and *Halomonas* sp. are potential flocculating agents in the treatment of industrial wastewater effluents. Up to 250% increases in biofloculant production were achieved by manipulation of the media composition. Increases in peptone and glycerol contents, up to 2 and 3%, respectively enhanced the biofloculants production. With the exception of *Herbaspirillum* sp. CH7, all isolates preferred higher yeast extract content for biofloculant production. Physicochemical analysis showed most of biofloculants produced appeared to contain high protein content, then carbohydrate. All biofloculants contained similar hexosamine and uronic acid contents in the range of 0.0115 to 0.0150 mM and 0.0054 to 0.0068 mM, respectively. Purified biofloculants produced under optimized conditions possessed higher flocculating activities (up to 14-fold increases) in comparison to those under the control conditions with the exception of that from *Herbaspirillum* sp. CH7. A decrease in biofloculant concentration from 10000 ppm to 1 ppm resulted in a significant increase in the removal percentages of Pb^{+2} , Zn^{+2} and Hg^{+2} with the optimal dosage of 1 to 10 ppm. Biofloculants in this study removed Cd^{+2} effectively only at 10000 ppm, but not at the lower concentrations. This Cd^{+2} -removing capacity of biofloculant could be further stimulated by an increase in temperature. The pH requirement for maximum flocculating activity seems to be varied for different strains of microorganisms.

Key words: Biofloculant, heavy metal removal, physicochemical properties, pH, temperature.

INTRODUCTION

Flocculation has been applied to a wide range of fields, such as dredging, wastewater treatment and fermentation. Due to their economic advantage and potency, synthetic organic and inorganic flocculants are widely used. However, serious environmental pollutions caused by these materials have been reported in the literature (Nakata and Kurane, 1999; Ahluwalia and Goyal, 2007; Chakraboti et al., 2008). Biofloculants as an alternative choice have gained more attention recently since they are environmentally friendly, biodegradable and nontoxic

(Salehizadeh and Shojaosadati, 2001).

Biofloculants have been used to flocculate inorganic solid suspensions (Natarajan and Das, 2003; Lu et al., 2005; Yim et al., 2007; Zhang et al., 2007), humic acids (Zouboulis et al., 2004), dyes solution (Deng et al., 2005; Gao et al., 2009; Liu et al., 2009), and others (Salehizadeh and Shojaosadati, 2003; Deng et al., 2003; Vijayalakshmi and Raichur, 2003; Lu et al., 2005; Aguilera et al., 2008). A number of studies have also demonstrated the potential for the use of biofloculants in heavy metal removal (Quintelas et al., 2008; Salehizadeh and Shojaosadati, 2003; Wu and Ye, 2007; Gong et al., 2008). Gong et al. (2008) demonstrated that biofloculant produced by *Serratia ficaria* was more effective in removing chemical oxygen demand and the color of pulp effluent than traditional chemical flocculants. Lin and Harichund (2011a) have previously demonstrated the ability of bacterial biofloculants simultaneously in

*Corresponding author. E-mail: linj@ukzn.ac.za. Tel: +27-31-2607407. Fax: +27-31-2607809.

Abbreviations: CPC, Cetyl pyridinium chloride; NRF, national research foundation

removing bacterial populations, heavy metals and turbidity from 3 industrial effluents.

A number of bioflocculant-producing bacteria have been discovered (Jiyoung et al., 2007). Wu and Ye (2007) conclude that microbial bioflocculant are largely composed of high molecular weight biopolymers such as polysaccharide flocculants (*Alcaligenes latus* B-18, *Alcaligenes cupus* KT201 and *Bacillus* sp. Dp-152), glycoprotein flocculants (*Arthrobacter* sp.) and protein flocculants (*Rhodococcus erythropolis* and *Bacillus* sp. Dp-152). Kumar et al. (2004) reported that the molecular weight and functional groups in the molecular chains of bioflocculant are important determinants for the flocculating activity. However, the high production costs and low yield incurred to produce bioflocculants have been the rate-limiting steps for their applications (Smith and Miettinen, 2006). Research is needed to identify and select new bioflocculant-producing microorganisms and to learn how to optimize fermentation conditions to enhance bioflocculant yields (He et al., 2002; Yang et al., 2007; Wang et al., 2007). The purpose of this present study was to investigate bioflocculant production by selected microorganisms. In addition, the physico-chemical properties of purified bioflocculants were characterized as well as their flocculating properties in removing heavy metals and solid suspension.

MATERIALS AND METHODS

Bacterial strains and media

Bioflocculant-producing strains of *Pseudomonas* sp. CH6, *Bacillus* sp. CH15, *Herbaspirillum* sp. CH7, and *Paenibacillus* sp. CH11 isolated previously from an industrial effluent sample containing heavy metals in a previous study (Lin and Harichund, 2011a) and *Halomonas* sp. obtained from the stock cultures of the Department of Microbiology at UKZN-Westville were used in this study. Pure cultures were maintained on YMPG agar containing 0.3% Yeast extract, 0.3% malt extract, 0.5% peptone, 1% D-glucose and 2% bacteriological agar at pH 7 (Nakata and Kurane, 1999). Stock cultures were subjected to cryopreservation for long-term storage. All assays were conducted in triplicate and the values were presented as mean \pm standard deviation.

Production of bacterial bioflocculants

Isolates were cultivated in 250 ml Erlenmeyer flasks containing 30 ml YMPG media for 20 h at 28°C at 220 rpm (Nakata and Kurane, 1999). Aliquots (0.7 ml) of cultivated bacterial isolates were then inoculated into a 250 ml Erlenmeyer flask containing 70 ml of production medium (0.5% yeast extract, 0.5% polypeptone, 2% ethanol, 1% glycerol, 0.05% K₂HPO₄, 0.05% MgSO₄.

7H₂O, 0.2% NaCl, and 0.2% CaCO₃) (Nakata and Kurane, 1999). The flasks were incubated for three days at 28°C at 220 rpm. To determine the optimal production of bioflocculants of each isolate, the composition of production medium was modified with various concentrations of carbon sources (ethanol and glycerol), nitrogen sources (peptone and yeast extract) and NaCl as shown in Table 1.

30 ml of the cell-free supernatant, obtained after centrifugation (6000 \times g for 15 min), was mixed with 60 ml of ethanol and the mixture was stored overnight at 4°C for precipitation of bioflocculants. Bioflocculant pellets were recovered by centrifugation at 6000 \times g for 15 min. Bioflocculant pellets were dried in a desiccator containing anhydrous cobalt chloride under reduced pressure at room temperature (~25°C) and the dry mass of bioflocculants was obtained. Recovered bacterial cells were re-suspended in 5 ml 0.9% NaCl and filtered using No. 1 filter paper (Whatman). Cells present on the filter paper were dried at 110°C in an oven. The cell mass was recorded after the weight of the dry cell did not change during consecutive measurements (Nakata and Kurane, 1999).

Purification of bacterial bioflocculants

Bioflocculants were further purified using the method of Xia et al. (2008). Dried bioflocculants were resuspended in sterile distilled water (50 ml) followed by the addition of 25 ml of 2% cetyl pyridinium chloride solution (CPC) with constant stirring for 2 h. The bioflocculant and CPC complex was obtained by centrifugation at 4000 \times g for 30 min, was resuspended in 50 ml of 0.5 M NaCl. Two volumes (100 ml) of cold ethanol were added to the mixture. Purified bioflocculants were obtained through centrifugal separation as described above. Precipitated bioflocculants were washed twice with 100% ethanol and the precipitate, vacuum dried in a desiccator as described above. After purification of the bacterial bioflocculants, the mass of the purified bioflocculant was measured. The bioflocculants were standardized to 10000 ppm as a stock solution and were used throughout the study.

Flocculating activity of purified bacterial bioflocculants

Flocculating activity of the bioflocculants was determined using the method of Nakata and Kurane (1999) using kaolin clay. One ml of bacterial bioflocculant (1000 ppm) was mixed with 9 ml of kaolin solution (5.5 g/l) and 100 μ l of CaCl₂·2H₂O (30 g/l). The mixture was vortexed for 30 s and rested for 5 min. The upper layer of the solution (3 ml) was then removed for further analyses. The flocculating activity was calculated based on OD_{550nm} as described by Kurane et al. (1994). The turbidity of

Table 1. Media composition for optimal bioflocculant production by selected bacterial isolates

Isolate	Media composition					Bioflocculant production		
	Ethanol	Glycerol	Yeast extract	Peptone	NaCl	Standard concentration (g/L)	Optimal concentration (g/L)	Ratio increase (%)
<i>Pseudomonas</i> sp. CH6	2%	3%	2%	1%	0.3%	0.715	1.43	200
<i>Herbaspirillum</i> sp. CH7	2%	2%	0.5%	1%	0.5%	0.571	1.00	175
<i>Paenibacillus</i> sp. CH11	2%	3%	2%	2%	0.4%	0.857	1.86	217
<i>Bacillus</i> sp. CH15	2%	3%	1%	2%	0.5%	0.571	1.42	250
<i>Halomonas</i> sp.	2%	2%	1%	1%	0.3%	1.714	3.57	208

Standard controlled conditions. 2% ethanol, 1% glycerol, 0.5% yeast extract, 0.5% peptone, 0.2% NaCl, 0.05% K₂HPO₄, 0.05% MgSO₄·7H₂O and 0.2% CaCO₃ @ 28°C, 220 rpm for 3 days.

samples (A) and control (B) was measured at 550 nm using a spectrophotometer (Varian Cary 50 UV-Visible Spectrophotometer). The control experiment was performed using 1 ml of water (Millipore Elix purification system, 17 mega Ω) without bioflocculant. The flocculating activity and removal rate was determined using the method of Nakata and Kurane (1999).

Heavy metal removal efficiency of purified bacterial bioflocculants

Heavy metal removal efficiency of bacterial bioflocculants was determined using the modified method of Nakata and Kurane (1999) as described by Wu and Ye (2007) using heavy metal solutions without kaolin clay. Different bioflocculant concentrations ranging from 1 ppm to 10000 ppm were suspended in the various concentrations of heavy metal solutions (0.5, 1 and 10 ppm, respectively) for 5 min. The heavy metal concentrations of the upper layer of the solution (3 ml), as described above, were also measured using ICP-OES.

Composition analysis of purified bacterial bioflocculants

The composition of each purified bioflocculant was determined using a variety of analyses. The carbohydrate, uronic acid and hexosamine contents of bioflocculants were determined using the phenol-sulfuric acid, carbozole assay and Morgan-Elson assays, respectively (Chaplin, 1994). The protein composition of bioflocculants was determined using the Folin-Lowry method (Plummer, 1978).

Effect of pH and temperature on heavy metal removal of purified bioflocculants

The effect of pH and temperature on heavy metal removal by bioflocculants was assayed under the specific

concentrations of heavy metal and bioflocculant for which the optimal removal has been achieved. A pH range of 3 to 9 was chosen to assay the effect of pH on removal of heavy metals. A temperature range of 4 to 45°C was selected to determine the effect of temperature on heavy metal removal using bioflocculants.

Statistical analysis

The experiments were conducted in triplicate and the results were expressed as mean \pm S.D. Paired *t*-tests, using SPSS version 15, were used to examine the statistical significance between different treatments. Probability was set at 0.05 as significance.

RESULTS

Optimization of bioflocculants productions

Experiments were conducted to optimize bioflocculant production, which involved the variation of carbon source concentration (ethanol and glycerol), nitrogen source concentration (yeast extract and peptone) and NaCl concentrations. Table 1 shows the media composition for the optimal bioflocculant production by selected bacterial isolates. By modifying carbon, nitrogen and NaCl compositions, bacterial bioflocculant production was increased by 175 to 250% (Table 1). The different bacterial isolates had different combination preferences for their optimal bioflocculant production. All isolates produced the highest amounts of bioflocculants using the same ethanol content (2%) as in the controlled conditions. Most of these isolates preferred higher contents of glycerol, yeast extract, peptone and NaCl in order to achieve a better production compared to the corresponding compositions under the controlled conditions. The highest amount of bioflocculant production prior to and after optimization was observed for *Halomonas* sp. (Table 1). No trend was observed between accumulated cell mass and bioflocculant productions (data not

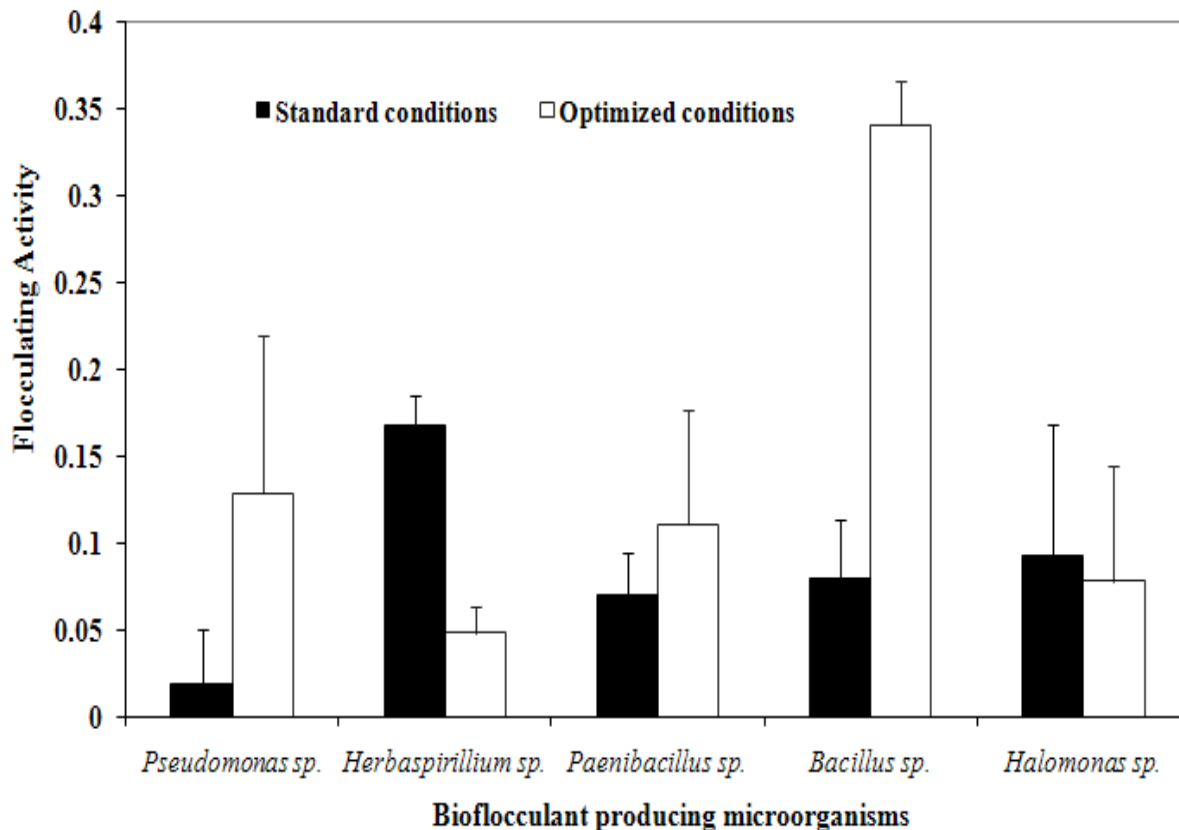


Figure 1. Flocculating activity of purified bacterial biofloculants (1000 ppm) obtained under the standard and optimized production conditions.

Table 2. Physicochemical analysis of partially purified bacterial biofloculants.

Isolate	Carbohydrate (mM)	Protein (mM)	Hexosamine (mM)	Uronic acid (mM)
<i>Pseudomonas sp.</i> CH6	0.1755	0.1775	0.0134	0.0057
<i>Herbaspirillum sp.</i> CH7	0.1995	0.1305	0.0115	0.0054
<i>Paenibacillus sp.</i> CH11	0.2741	1.1012	0.0145	0.0057
<i>Bacillus sp.</i> CH15	0.3446	1.5780	0.0150	0.0060
<i>Halomonas sp.</i>	0.0054	0.1775	0.0119	0.0068

shown).

Flocculating activity of crude and purified biofloculants

Crude bacterial biofloculants produced under the controlled and optimal conditions were further purified using 2% CPC. The flocculating activities of purified biofloculants from both conditions are shown in Figure 1. Flocculating activities of biofloculants produced by *Pseudomonas sp.* CH6 and *Bacillus sp.* CH15 were increased, but not significant ($p > 0.05$), up by 1400 and 350%, respectively while that of *Paenibacillus sp.* CH11 increased by 140% (Figure 1). However, the flocculating

activity decreased from 0.168 to 0.049 of the purified biofloculant of *Herbaspirillum sp.* under controlled conditions to those for optimal conditions (Figure 1).

Physicochemical analysis of purified bacterial biofloculants

Physicochemical analyses of purified biofloculants were conducted to determine the composition of the bacterial biofloculants (Table 2). Biofloculants produced by *Paenibacillus sp.* CH11, *Bacillus sp.* CH15 and *Halomonas sp.* comprised of protein as the major component while that of *Pseudomonas sp.* CH6 contained carbohydrate and protein in almost equal

amounts. *Herbaspirillum* sp. CH7 differed in that it possessed more carbohydrate content. All bacterial bioflocculants in this study had similar levels of uronic acid and hexosamine composition. Bioflocculants produced were found to have the uronic acid and hexosamine contents in the range of 0.0115 to 0.0150 mM and 0.0054-0.0068 mM, respectively (Table 2).

Effect of purified bioflocculant concentrations in heavy metal removal

The effect of purified bioflocculant concentrations on heavy metal removal is presented in Figure 2. In general, decreases in bioflocculant concentration resulted in a significant increase in the percentages of metal removal except in the case of cadmium. As shown in Figure 2a, bioflocculant produced by *Pseudomonas* sp. dramatically increased the efficiency of Zn^{+2} removals from 24 to 46% as the concentration dropped from 10000 to 1000 ppm, respectively and slowly increased the efficiency as the concentration continued to decrease. The same trends were observed when removing Pb^{+2} and Hg^{+2} using the same bioflocculant (Figure 2b and 2c) as well as the bioflocculants produced by other bioflocculant-producing bacteria in this study. Metal-removing capacity of the bioflocculant produced by *Halomonas* sp. was more sensitive to changes in concentrations, and Hg^{+2} in particular (Figure 2c). Eighty percentages of mercury (1 ppm) was efficiently removed by one ppm of the bioflocculant from *Herbaspirillum* sp as shown in Figure 2c. However, a reverse trend was observed when these bioflocculants were used to remove Cd^{+2} . The Cd^{+2} removal efficiency of *Pseudomonas*' bioflocculant was significantly decreased from 66 to 19% while that of the *Herbaspirillum* sp. bioflocculant decreased from 66 to 10% when the bioflocculant concentration was reduced from 10000 to 1000 ppm, respectively (Figure 2d). Bioflocculant produced by *Halomonas* sp. had a lower Cd^{+2} removal capacity compared to other isolates in this study.

Effects of pH and temperature on heavy metal removal using bacterial bioflocculants

The effect of pH on heavy metal removal efficiency of purified bioflocculants at the optimal removal concentrations is demonstrated in Figure 3. In general, the bacterial bioflocculants in this study had better heavy-metal removing capacity for all metals tested at pH 5 and 7 (Figure 3). At pH 3, the efficiency of all bioflocculants for binding Zn^{+2} decreased from 40 and 50% to 10 and 20%. The efficiency in binding Zn was dropped from around 46% to 10 and 20% when the pH was adjusted to 3 for all bioflocculants. At pH 9 and 10 and 10000 ppm, binding efficiencies of Cd^{+2} and Hg^{+2} decreased in pH 9

for all tested bioflocculants with the exception of *Bacillus* sp. (Figure 3c and 3d). Bioflocculant produced by *Paenibacillus* sp. removed only 5% of Cd^{+2} at pH9 compared to 87% at pH3. However, bioflocculant produced by *Bacillus* sp. was capable of removing 64% of Cd^{+2} at pH7, compared to 4.4% at pH 3. Changes in pH had less impact in removing Pb^{+2} using bioflocculants (Figure 3b). The results in Figure 4 showed the percentage of heavy metal removal at different temperatures (4 and 45°C) by *Paenibacillus* sp. CH11 bioflocculant at the optimal removal concentrations observed in Figure 2. No trend was observed in the relationship between temperature and heavy metal-removing ability of this bioflocculant at low concentrations in the case of Zn^{+2} , Pb^{+2} and Hg^{+2} . Altering the temperature within the 4 and 45°C range did not produce any significant changes on flocculating activities of bioflocculants in this study. However, there was an increase in the Cd^{+2} removal from 20% up to 93% as the temperature increased from 4 to 45°C using a high concentration (10000 ppm) of bioflocculant (Figure 4). Similar results were obtained with other bioflocculants in this study (data not shown).

DISCUSSION

Bioflocculants have been actively studied as an alternative choice in various industrial waste treatments recently (Aguilera et al., 2008; Gao et al., 2009; Liu et al., 2009; Yim et al., 2007) including heavy metal removal (Iyer et al., 2005; Wu and Ye, 2007; Gong et al., 2008; Lin and Harichund, 2011a). Due to their extensive capacity for metals, bioflocculants are recommended as surface-active agents for the removal of heavy metals (Morillo et al., 2006). However, bioflocculant production costs are high compared to conventional approaches (Lian et al., 2008). Therefore, optimal bioflocculant production and high flocculating activity of bioflocculant at a lower dosage are crucial for applications.

Optimization of bioflocculants productions

Elements such as culture medium and culture conditions are known to affect bioflocculant production (Noghabi et al., 2007; Xia et al., 2008). Previous studies on optimization of production yield revealed optimum bioflocculant production occurred in complex media such as yeast extract and peptone (Liu et al., 2010; Nakata and Kurane, 1999; Xia et al., 2008; Zheng et al., 2008). Decreased bioflocculant production of *Bacillus licheniformis* due to the presence of peptone in the production medium has also been reported (Shih et al., 2001). In the current study, carbon source (ethanol and glycerol), nitrogen source (yeast extract and peptone) and NaCl concentrations were varied and the effect of these

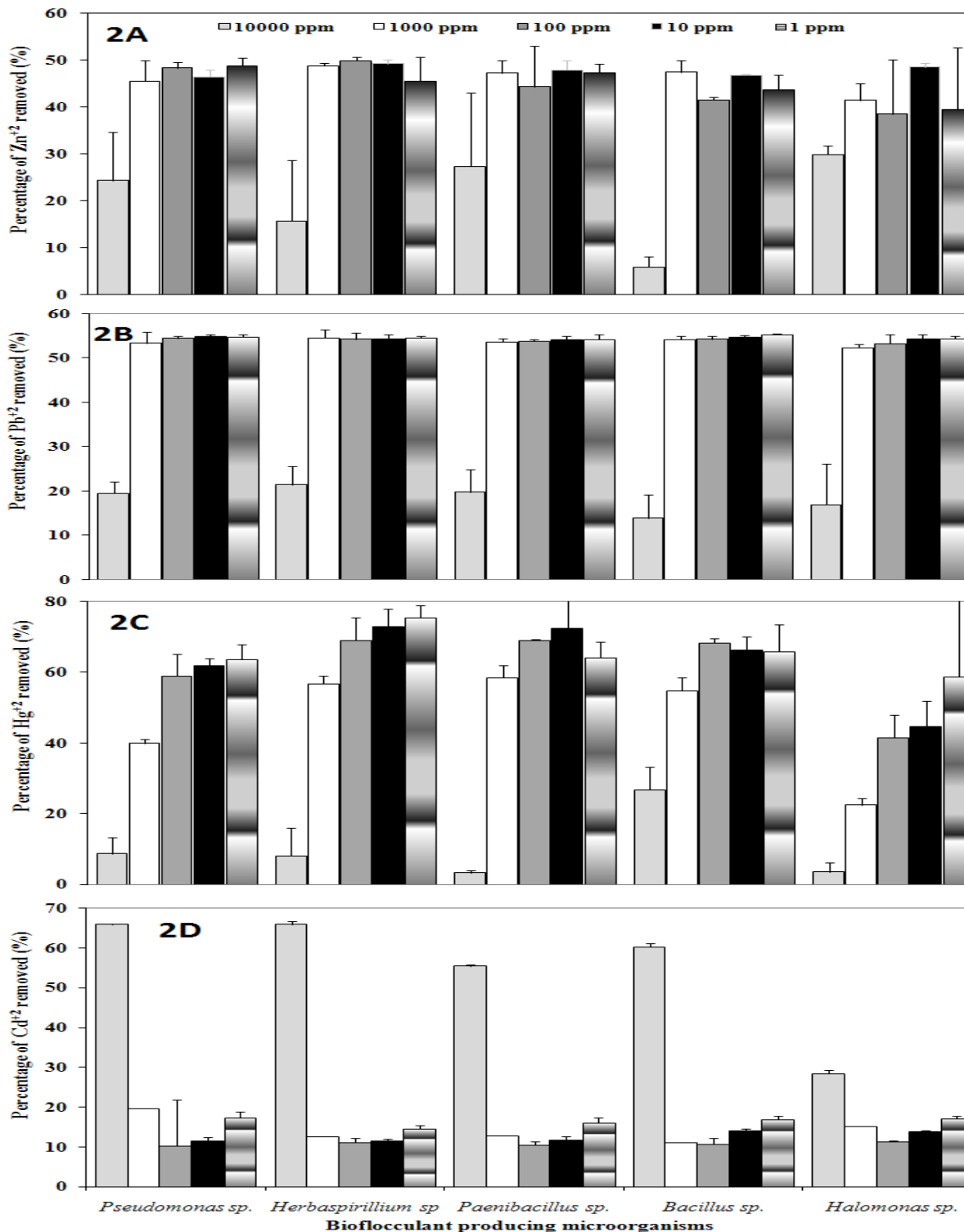


Figure 2. Percentage removal of heavy metal (A, zinc; B, lead; C, mercury; D, cadmium at 1 ppm) through biofloculation using a range of biofloculant concentrations (1 to 10000 ppm).

variations on biofloculant production was assessed. Our results further support previous studies, since organic nitrogen such as yeast extract and peptone (He et al.,

2004; Xia et al., 2008, Zheng et al., 2008) were better sources than inorganic nitrogen for biofloculant production. Increasing in the yeast extract and peptone

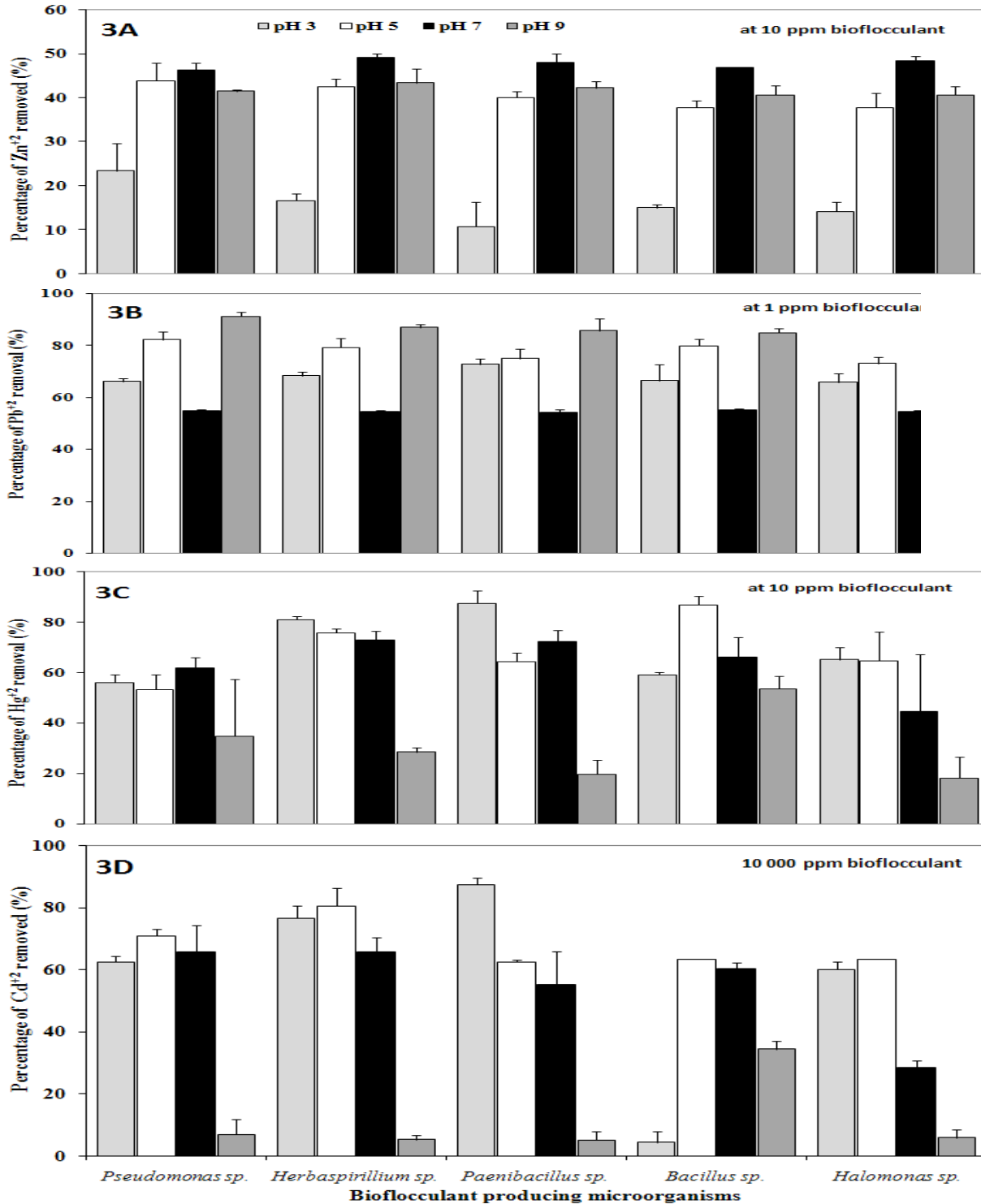


Figure 3. Percentage removal of heavy metal at 1 ppm (A, Zn²⁺; B, Pb²⁺; C, Hg²⁺; D, Cd²⁺) at different pH using bacterial bioflocculants at the optimal removal concentrations.

contents from the standard compositions recommended by Kurane et al. (1994) enhanced the bioflocculant productions by all bacterial isolates (Table 1) except for *Herbaspirillum sp.* CH7. Similar to the reports by

Noghabi et al. (2007) and Freitas et al. (2009), high glycerol concentration in the medium also stimulated the bioflocculation production. Up to 250% increases in bioflocculant production was obtained in this study. He et

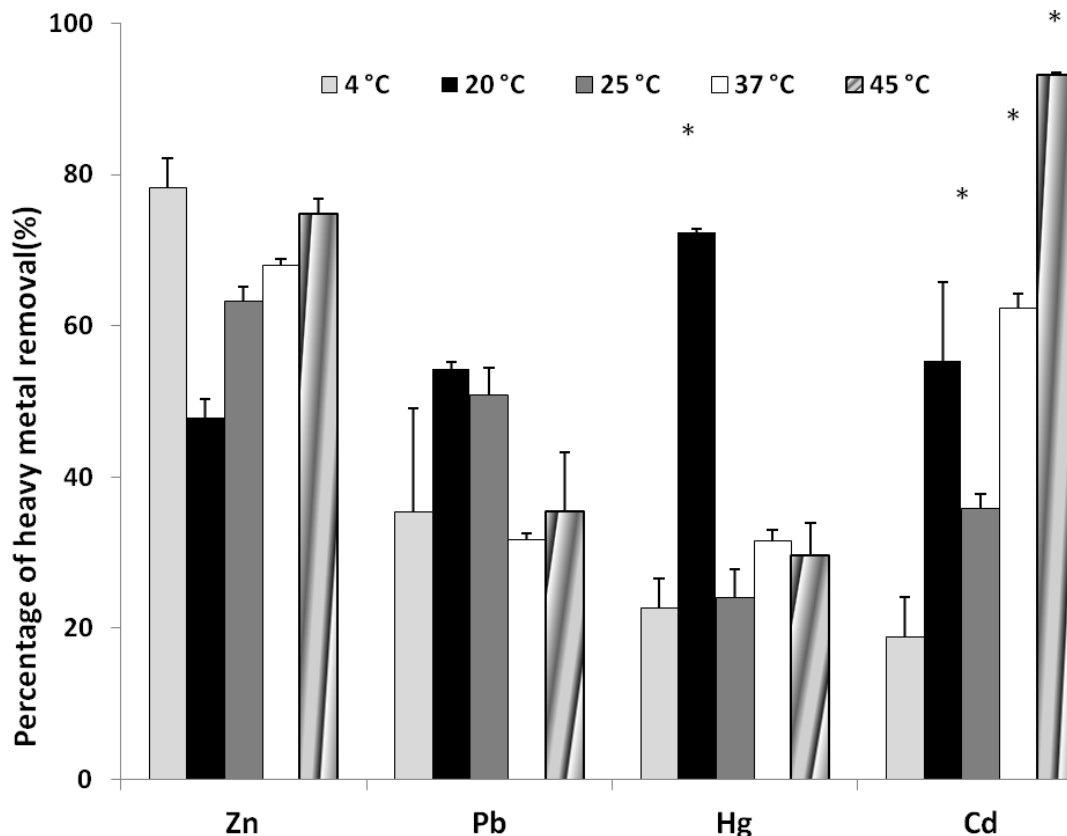


Figure 4. Percentage removal of heavy metal (Zn^{+2} ; Pb^{+2} ; Hg^{+2} ; Cd^{+2}) at 1 ppm at different temperature using *Paenibacillus* sp. CH11 bioflocculant at the optimal removal concentrations (Figure 2). *: significant, $p < 0.05$.

al. (2009) reported that edible glucose was found to be one of the significant factors affecting bioflocculant production by *Halomonas* sp.

Flocculating activity of bioflocculants

Although, an increase in flocculation was observed using purified bioflocculants produced by *Pseudomonas* sp. CH6, *Paenibacillus* sp. CH11 and *Bacillus* sp. CH15 under the optimal conditions compared to those from the controlled conditions, there were not significant ($p > 0.05$). A significant decrease in flocculating activity was observed for the bioflocculant from *Herbaspirillum* sp. CH7. A direct correlation between increased bioflocculant production and increased/decreased flocculating activity was not evident. Wu et al. (2005) reported that different microorganisms produce bioflocculants using different compositions resulting in different flocculating activities. Bacteria and fungi, belonging to different species, have been reported to produce different EPS types that possess different characteristics and function optimally at various media concentrations (Subramanian et al., 2007). Changes to media composition may have

positively or negatively affected the contents of functional groups such as carboxyl and hydroxyl groups of bioflocculants produced resulting in increased (*Pseudomonas* sp. CH6, *Paenibacillus* sp. CH11 and *Bacillus* sp. CH15) or decreased (*Herbaspirillum* sp. CH7) flocculation. Another possibility that arises to explain fluctuations in flocculating activity is that cations ($CaCl_2$) may affect flocculating activity by neutralizing and stabilizing the residual negative charge of functional groups by forming bridges between particles (Wu and Ye, 2007). Our previous study (Lin and Harichund, unpublished results) showed that the bioflocculants used in this study had better flocculating activities when other heavy metal (Pb^{+2} , Zn^{+2} or Hg^{+2}) were used instead of Ca^{+2} . Several groups also found that Ca^{+2} had little effect on the flocculating activity of produced bioflocculant (Sheng et al., 2006; Yim et al., 2007; Zheng et al., 2008).

Physicochemical analysis of purified bacterial bioflocculants

Bioflocculants have been classified into four types based on their main chemical constituent, that is, polysac-

charide (Wu and Ye, 2007, 2004; Xia et al., 2008; Yu et al., 2009; Zheng et al., 2008), protein (Liu et al., 2010; Park and Novak, 2007), lipid (Kurane et al., 1994; Guiband et al., 2005) and/or nucleic acid/nucleoprotein bioflocculants (Wang et al., 1994). Polysaccharides and proteins offer numerous active sites capable of binding metal ions (Saiffudin and Raziah, 2007). Most of the bioflocculants in the present study can be considered as protein bioflocculants since their main components were proteins, with the exception of the *Herbaspirillum* sp. CH7 bioflocculant which was polysaccharide. All the bioflocculants contained similar contents of hexosamine and uronic acid. Bioflocculants produced by *Halomonas eurihalina* and by *Bacillus subtilis* DYU1 contained 11.1% (Bejar et al., 1998) and 2.7% uronic acid (Wu and Ye, 2007), respectively. All these compositions are involved in the flocculating ability. Microbial bioflocculants involved multiple adsorption points/structures for their flocculating abilities so that bioflocculants are capable of forming the floc with different varieties of organic, inorganic particles as well as microorganisms from the environments as demonstrated by our previous study (Lin and Harichund, 2011b) and others (Gong et al., 2008; Quintelas et al., 2008).

Effect of bioflocculant concentrations in removal of heavy metals

Bioflocculants in this study were effective as heavy metal flocculating agents with low dosage requirement as the optimal bioflocculant concentration in removing Zn^{+2} , Pb^{+2} and Hg^{+2} in the range of 1 to 10 ppm (Figure 2). Yim et al. (2007) demonstrated that sedimentation of flocculated particles was inhibited by the viscosity generated at high concentrations of bioflocculants. Over-addition of negatively charged bioflocculant caused the repulsion of each other and poor stability. The relationship between bioflocculant dosage and flocculating rate was similar to that of the bioflocculants produced by other pure strains (Zheng et al., 2008; Gong et al., 2008; Yim et al., 2007; Liu et al., 2009). The optimal bioflocculant dosage was often reported in the similar range (He et al., 2009; Yim et al., 2007; Zheng et al., 2008). The amino and carboxyl functional groups of bioflocculant can form a floc with heavy metals by neutralizing and stabilizing the residual charge as the binding distance is shortened. The physical and chemical properties of the metals, the availability of appropriate binding sites present to metal, as well as the tertiary structure of bioflocculant may all contribute to metal-binding interactions (Kachlany et al., 2001). Surprisingly, high concentrations (10000 ppm) of bioflocculants were needed to effectively remove Cd^{+2} from the medium (Figure 2). It is the first report demonstrating the binding of heavy metals and bioflocculants at this high dosage. It is generally believed that bioflocculants have a low metal binding capacity or

flocculating activity at higher dosage due to the repulsion and low stability (Zheng et al., 2008; Yim et al., 2007; Liu et al., 2009). All the bioflocculant studies in the literature conducted the experiments using less than 2000 ppm of bioflocculants. It is possible that other metal-binding motifs might exist for the bioflocculant. In biological systems, amino acids, like Cys residues of the metal binding CXXC motif, are responsible of Cd^{+2} tolerance (Banci et al., 2006; Wu et al., 2006). Bioflocculants in this study comprised of high protein contents. It will be interesting to study the Cd^{+2} binding motifs of these bioflocculants. However, the application to remove Cd^{+2} using bioflocculants at the high dosages may be limited.

Effect of pH on heavy metal removal using bacterial bioflocculants

The pH of the suspension had an extreme effect on the function of flocculating agent including bioflocculant because the surface charge of the dispersed phase changed according to various pH values (Sheng et al., 2006). Cations such as heavy metal ions can neutralize negatively-charged functional groups of bioflocculant molecules and weaken the static repulsive force, and promote floc formation (Li et al., 2008; He et al., 2010). These results could be attributed to double layer compression effects of divalent cation (Zheng et al., 2008; Li et al., 2008, 2009). Bioflocculants may either enhance or decrease heavy metal removal due to the change of pH (Kaewchai and Prasertsan, 2002; Kiran and Kaushik, 2008; Panicker et al., 2006). Gupta et al. (2000) demonstrated that binding of metals to bioflocculants may be reversed by pH adjustments indicating that the type of binding is pH-dependent. Depending on the functional groups of bioflocculants that are responsible of binding the heavy metals, microbial bioflocculants were likely to absorb hydrogen ions (H^+) at low pH, which might weaken the forming of complexes between bioflocculant and heavy metals. By the same token, hydroxide ions (OH^-) interfered with the combination of the flocculant molecules and heavy metals at high pH, resulting in lower flocculating activity. As observed previously by other researchers (Li et al., 2008; Liu et al., 2009; He et al., 2010), flocculating effect of most bioflocculants appeared to be the strongest at neutral pH values in general. Bioflocculants produced by *Bacillus* sp. F19 and *Gyrodinium impudicum* KG03 had the highest flocculating activity at pH 2.0 and pH 4.0, respectively (Zheng et al., 2008; Yim et al., 2007). Bioflocculants from *Agrobacterium* sp. M-503 had the optimal flocculating activity between pH 8-12 (Li et al., 2010). The pH requirement for maximum flocculating activity seems to be varied for different strains of microorganisms (Xia et al., 2008).

Temperature factor

In biological wastewater treatment processes, tempera-

ture fluctuations occur, due to seasonal variation. This temperature shift results in alteration in the efficiency of the treatment process. Flocculation has been found to be affected by temperature shifts (Sagastume and Allen, 2003). Pure protein bioflocculants, where the amino and carboxyl groups are the effective groups for bioflocculation (Kurane et al., 1994), usually not heat-stable as protein destroyed upon heating (Takeda et al., 1991). If the major component of a bioflocculant is glycoprotein that has many functional groups (Kurane et al., 1991), its stability will depend on the relative content of protein and polysaccharide. The results obtained in the present study indicated no correlation between the metal removal patterns by bioflocculants and specific temperatures using low concentration of bioflocculants. At higher dosages, bioflocculants removed Cd^{+2} effectively and this was increased with an increase in temperature under the optimal conditions. At high concentrations of bioflocculants, the binding of Cd^{+2} and bioflocculants might neutralize negatively-charged functional groups of bioflocculant molecules partially. Further increases in temperature might weaken the static repulsive force and promote floc formation (Li et al., 2008; He et al., 2010). Further studies are needed.

Conclusions

The present study has shown that up to 250% increases in bioflocculant productions were achieved by manipulation of media components. Purified bioflocculants produced under the optimal conditions also possessed higher flocculating activities (up to 14-fold increases) compared with those produced the controlled conditions. The results also showed that the optimal dosages of purified bioflocculants in removing Pb^{+2} , Zn^{+2} and Hg^{+2} was in the range of 1 to 10 ppm. Surprisingly, high dosages of bioflocculants were needed in Cd^{+2} removal, which was also stimulated by an increase in temperature. The pH requirement for maximum heavy metal removal activity seems to be varied for different strains of microorganisms.

ACKNOWLEDGEMENTS

This research was funded by the National Research Foundation (NRF) of South Africa. The authors would like to thank Dr. H. Chenia in the same Department for editing the manuscript. The second author is grateful to the NRF for the award of Masters bursary.

REFERENCES

Aguilera M, Quesada MT, Aguila VGD, Morillo JA, Rivadeneyra MA, Cormenzana AR, Sanchez MM (2008). Characterization of *Paenibacillus jamilae* strains that produce exopolysaccharide during growth on and detoxification of olive mill wastewaters.

- Bioresour. Technol. 99: 5640-5644.
- Ahluwalia SS, Goyal D (2007). Microbial and plant derived biomass for removal of heavy metals from wastewater. *Bioresour. Technol.* 98: 2243-2257.
- Banci L, Bertini I, Ciofi-Baffoni S, Su XC, Miras R, Bal N, Mintz E, Catty P, Shokes JE, Scott RA (2006). Structural Basis for Metal Binding Specificity: the N-terminal Cadmium Binding Domain of the P1-type ATPase CadA. *J. Mol. Biol.* 356: 638-650.
- Bejar V, Llamas I, Calvo C, Quesada E (1998). Characterization of exopolysaccharides produced by 19 halophilic strains of the species *Halomonas eurihalina*. *J. Biotechnol.* 61: 135-141.
- Chaplin MF (1994). Monosaccharides, pp 2-6. In M.F. Chaplin and J.F. Kennedy, *Carbohydrate Analysis* (2nd ed.), Oxford University Press, New York.
- Deng SB, Bai RB, Hu XM, Luo Q (2003). Characteristics of a bioflocculant produced by *Bacillus mucilaginosus* and its use in starch wastewater treatment. *Appl. Microbiol. Biotechnol.* 60: 588-593.
- Deng SB, Yu G, Ting YP (2005). Production of a bioflocculant by *Aspergillus parasiticus* and its application in dye removal. *Colloids Surf. B: Biointerfaces* 44: 179-186.
- Freitas F, Alves VD, Pais J, Costa N, Oliveira C, Mafra L, Hilliou L, Oliveira R, Reis MAM (2009). Characterization of an extracellular polysaccharide produced by a *Pseudomonas* strain grown on glycerol. *Bioresour. Technol.* 100: 859-865.
- Gao Q, Zhu X-H, Mu J, Zhang Y, Dong X-W (2009). Using *Ruditapes philippinarum* conglutination mud to produce bioflocculant and its applications in wastewater treatment. *Bioresour. Technol.* 100: 4996-5001.
- Gong WX, Wang SG, Sun XF, Liu XW, Yue QY, Gao BY (2008). Bioflocculant production by culture of *Serratia ficaria* and its application in wastewater treatment *Bioresour. Technol.* 99: 4668-4674.
- Guiband, G., S. Comte., F. Bordas., S. Dupuy., and M. Baudau. 2005. Comparison of the complexation potential of extracellular polymeric substances (EPS), extracted from activated sludges and produced by pure bacteria strains, for cadmium, lead and nickel. *Chemosphere* 59: 629-638.
- Gupta R, Ahuja P, Khan S, Saxena RK, Mohapatra H (2000). Microbial biosorbents: meeting challenges of heavy metal pollution in aqueous solutions. *Curr. Sci.* 78: 967-973.
- He N, Li Y, Chen J, Lun SY (2002). Identification of a novel bioflocculant from a newly isolated *Corynebacterium glutamicum*. *Biochem. Eng. J.* 11: 137-148.
- He N, Li Y, Chen J (2004). Production of a novel polygalacturonic acid bioflocculant REA-11 by *Corynebacterium glutamicum*. *Bioresour. Technol.* 94, 99-105.
- He J, Zhen QW, Qiu N, Liu ZD, Wang BJ, Shao ZZ, Yu ZJ (2009). Medium optimization for the production of a novel bioflocculant from *Halomonas* sp. V3a0 using response surface methodology. *Bioresour. Technol.* 100: 5922-5927.
- He J, Zou J, Shao Z, Zhang J, Liu Z, Yu Z (2010). Characteristics and flocculating mechanism of a novel bioflocculant HBF-3 produced by deep-sea bacterium mutant *Halomonas* sp. V3a'. *World J Microbiol Biotechnol* 26:1135-1141.
- Iyer A, Mody K, Jha B (2005). Biosorption of heavy metals by marine bacterium. *Marine Pollution Bulletin* 50: 340-346.
- Jiyoung Y, Lee YH, Jeon SM, Jung UJ, Lee MK, Jung YM, Choi MS (2007). Supplementation of a novel microbial biopolymer, PGB1, from new *Enterobacter* sp. BL-2 delays the deterioration of type 2 diabetic mice. *J. Microbiol. Biotechnol.* 17: 1983-1990.
- Kachlany SC, Lavery SB, Kim JS, Reuhs BL, Lion LW, Ghiorse WC (2001). Structure and carbohydrate analysis of the exopolysaccharide capsule of *Pseudomonas putida* G7. *J. Environ. Microbiol.* 3: 774-784.
- Kaewchai S, Prasertan P (2002). Biosorption of heavy metal by thermotolerant polymer-producing bacterial cells and the bioflocculant. *Songklanakarin J. Sci. Technol.* 24: 421-430.
- Kiran B, Kaushik A (2008). Chromium binding capacity of *Lyngbya putealis* exopolysaccharides. *Biochem. Eng. J.* 38: 47-54.
- Kumar CG, Joo HS, Choi JW, Koo YM, Yoon CS (2004). Purification and characterization of an extracellular polysaccharide from

- haloalkalophilic *Bacillus* sp. I-450. *Enzyme Microb. Technol.* 34: 673-681.
- Kurane R, Hatamochi K, Kakuno T, Kiyohara M, Kawaguchi K, Mizono Y, Hirano M, Taniguchi Y (1994). Purification and characterization of liquid bioflocculant produced by *Rhodococcus erythropolis*. *Biosci. Biotechnol. Biochem.* 58, 1977–1982.
- Kurane R, Hatamochi K, Kakuno T, Kiyohara M, Kawaguchi K, Mizono Y, Hirano M, Taniguchi Y (1991). Microbial flocculation of waste liquid and oil emulsion by a bioflocculant from *Alcaligenes latus*. *Agric Biol Chem* 55:1127–1129.
- Li Q, Liu HL, Qi QS, Wang FS, Zhang YZ (2010). Isolation and characterization of temperature and alkaline stable bioflocculant from *Agrobacterium* sp. M-503. *New Biotechnol.* 27 (6): 789-794.
- Li WW, Zhou WZ, Zhang YZ, Wang J, Zhu XB (2008). Flocculation behavior and mechanism of an exopolysaccharide from the deep sea psychrophilic bacterium *Pseudoalteromonas* sp. SM9913. *Bioresour. Technol* 99:6893–6899.
- Li Z, Chen RW, Lei HY, Shan Z, Bai T, Yu Q, Li HL (2009). Characterization and flocculating properties of a novel bioflocculant produced by *Bacillus circulans*. *World J Microb. Biotechnol.* 25:745–752.
- Lian B., Chen Y, Zhao J, Teng HH, Zhu L, Yuan S (2008). Microbial flocculation by *Bacillus mucilaginosus*: Applications and mechanisms. *Bioresour. Technol.* 99: 4825–4831.
- Lin J, Harichund C (2011a) Isolation and characterization of heavy metal removing bacterial bioflocculants. *Afri. J. of Microbiol. Res.* 5(6): 599-607.
- Lin J, Harichund C (2011b) Industrial effluent treatments using heavy-metal removing bacterial bioflocculants, *Water SA.* 37(2) 265-270.
- Liu WJ, Wang K, Li BZ, Yuan HL, Yang JS (2010). Production and characterization of an intracellular bioflocculant by *Chryseobacterium daeguense* W6 cultured in low nutrition medium. *Bioresour. Technol.* 101: 1044–1048.
- Liu WJ, Yuan HL, Yang JS, Li BZ (2009). Characterization of bioflocculants from biologically aerated filter backwashed sludge and its application in dyeing wastewater treatment. *Bioresour. Technol.* 100: 2629–2632.
- Lu WY, Zhang T, Zhang DY, Li CH, Wen JP, Du LX (2005). A novel bioflocculant produced by *Enterobacter aerogenes* and its use in defecating the trona suspension. *Biochem. Eng. J.* 27 1–7.
- Morillo JA, Aguilera M, Cormenzana AR, Sanchez MM (2006). Production of a metal-binding exopolysaccharide by *Paenibacillus jamilae* using two phase olive mill waste as fermentation substrate. *Curr. Microbiol.* 53: 189-193.
- Nakata K, Kurane R (1999). Production of an extracellular polysaccharide bioflocculant by *Klebsiella pneumoniae*. *Biosci. Biotechnol. Biochem.* 63: 2064-2068.
- Natarajan KA, Das A (2003). Surface chemical studies on 'Acidithiobacillus' group of bacteria with reference to mineral flocculation. *Int. J. Miner. Process* 72: 189-198.
- Noghabi KA, Zahiri HS, Yoon SC (2007). The production of cold-induced extracellular biopolymer by *Pseudomonas fluorescens* BM07 under various growth conditions and its role in heavy metal absorption. *Proc. Biochem. J.* 42: 847-855.
- Panicker G, Aislabie J, Bej AK (2006). Analysis of aggregative behavior of *Pseudomonas* sp. 30-3 isolated from Antarctic soil. *J. Soil Biol. Biochem.* 38: 3152-3157.
- Plummer DT (1978). *An Introduction to Practical Biochemistry*. 2nd Ed., pp 158-159. McGraw-Hill Press London.
- Quintelas C, Fernandes B, Castro J, Figueiredo H, Tavares T (2008). Biosorption of Cr (VI) by a *Bacillus coagulans* biofilm supported on granular activated carbon (GAC). *Chem. Eng. J.* 136: 195-203.
- Sagastume FM, Allen DG (2003). Effects of temperature transient conditions on aerobic biological treatment of wastewater. *Wat. Res.* 37: 3590-3601.
- Salehizadeh H, Shojaosadati SA (2001). Extracellular biopolymeric flocculants: Recent trends and biotechnological importance. *J. Biotechnol. Adv.* 19: 371-385.
- Salehizadeh H, Shojaosadati SH (2003). Removal of metal ions from aqueous solution by polysaccharide produced from *Bacillus firmus*. *Wat. Res.* 37: 4231–4235.
- Sheng YL, Zhang Q, She YR, Li CB, Wang HJ (2006). Screening and flocculating properties of bioflocculant-producing microorganisms. *Journal of University of Science and Technology Beijing* 13 (4): 289-292.
- Shih IL, Van YT, Yeh LC, Lin HG, Chang YN (2001). Production of a biopolymer flocculant from *Bacillus licheniformis* and its flocculation properties. *Bioresour. Technol.* 78: 267-272.
- Smith RW, Miettinen M (2006). Microorganisms in flotation and flocculation: Future technology or laboratory curiosity? *Miner. Eng.* 19: 548–553.
- Subramaniam SB, Yan S, Tyagi RD, Surampalli RY (2007). Characterization of extracellular polymeric substances (EPS) extracted from both sludge and pure bacterial strains isolated from wastewater sludge for sludge dewatering. *Wat. Res.* 12: 1-7.
- Takeda M, Kurane R, Koizumi JI, Nakamura I (1991). A protein bioflocculant produced by *Rhodococcus erythropolis*. *Agric. Biol. Chem.* 5:2663-2664.
- Vijayalakshmi SP, Raichur AM (2003). The utility of *Bacillus subtilis* as a bioflocculant for fine coal. *Colloids Surfaces B* 29: 265-75.
- Wang S-G, Gong W-X, Liu X-W, Tian L, Yue Q-Y, Gao B-Y (2007). Production of a novel bioflocculant by culture of *Klebsiella mobilis* using dairy wastewater. *Biochem. Eng. J.* 36: 81–86.
- Wang Z, Wang K, Xie Y (1995). Bioflocculant-producing microorganisms. *Acta Microbiol. Sin.* 35, 121-129.
- Wu CC, Gardarin A, Martel A, Mintz E, Guillain F, Catty P (2006). The Cadmium Transport Sites of CadA, the Cd²⁺-ATPase from *Listeria monocytogenes*. *J. Biol. Chem.* 281, 29533-29541.
- Wu JY, Ye HF (2007). Characterization and flocculating properties of an extracellular biopolymer produced from a *Bacillus subtilis* DYU1 isolate. *Proc. Biochem.* 42: 1114–1123
- Xia SQ, Zhang ZQ, Wang XJ, Yang A, Chen L, Zhao JF, Leonard D, Jaffrezic-Renault N (2008). Production and characterization of a bioflocculant by *Proteus mirabilis* TJ-1. *Bioresour. Technol.* 99: 6520-6527.
- Yang YN, Ren N, Xue JM, Yang J, Rong BL (2007). Mutation effect of MeV protons on bioflocculant bacteria *Bacillus cereus*. *Nucl. Instrum. Methods B* 262: 220–224.
- Yim JH, Kim SJ, Ahn SH, Lee HK (2007). Characterization of a novel bioflocculant, p-KG03, from a marine dinoflagellate, *Gyrodinium impudicum* KG03. *Bioresour. Technol.* 98: 361-367.
- Yu GH, He PJ, Shao LM (2009). Characteristics of extracellular polymeric substances (EPS) fractions from excess sludges and their effects on bioflocculability. *Bioresour. Technol.* 100: 3193-3198.
- Zhang ZQ, Lin B, Xia SQ, Wang XJ, Yang AM (2007). Production and application of a novel bioflocculant by multiple-microorganism consortia using brewery wastewater as carbon source. *Environ. Sci.* 19: 667–673.
- Zheng Y, Ye ZL, Fang XL, Li YH, Cai WM (2008). Production and characteristics of a bioflocculant produced by *Bacillus* sp F19. *Bioresour. Technol.* 99:7686-7691.
- Zouboulis AI, Chai XL and Katsoyiannis I (2004). The application of bioflocculant for the removal of humic acids from stabilized landfill leachates. *J. Environ. Manage.* 70: 35-41.